
This is the **accepted version** of the article:

Bes, Miquel; ; Sayol i Altarriba, Ferran; [et al.]. On the influence of water conductivity, Ph and climate on bryophyte assemblages in Catalan semi-natural springs. DOI 10.1080/03736687.2018.1446484

This version is available at <https://ddd.uab.cat/record/203572>

under the terms of the  **CC BY** COPYRIGHT license

1 **On the influence of water conductivity, pH and climate**
2 **on bryophyte assemblages in Catalan semi-natural**
3 **springs**

4 **Authors:** M. Bes¹, J. Corbera¹, F. Sayol², G. Bagaria², M. Jover³, C. Preece^{2,4}, F.
5 Sabater^{1,4,5}, A. Viza^{1,5}, M. Fernández-Martínez^{*1,6}

6 **Affiliations:**

7 ¹ ICHN, Delegació de la Serralada Litoral Central, Mataró, Catalonia, Spain

8 ² CREA, Cerdanyola del Vallès, 08193 Barcelona, Catalonia, Spain

9 ³ Department of Environmental Sciences, Faculty of Sciences, University of Girona,
10 Girona, Catalonia, Spain

11 ⁴ CSIC, Global Ecology Unit, CREA-CSIC-UAB, Cerdanyola del Vallès, 08193
12 Barcelona, Catalonia, Spain

13 ⁵ Department of Ecology, University of Barcelona, Barcelona, Catalonia, Spain

14 ⁶ Centre of Excellence PLECO (Plant and Vegetation Ecology), Department of Biology,
15 University of Antwerp, 2610 Wilrijk, Belgium.

16 ***Corresponding author**

17 Marcos Fernández-Martínez

18 E-mail address: marcos.fernandez-martinez@uantwerpen.be

19 **Manuscript length:** 3540 words

20 **Keywords:** Bryophyte ecology, Ecological niche, Fountains, Liverworts, Mosses,
21 Tolerance ranges

22 **Abstract**

23 Bryophytes are some of the most sensitive biological indicators of environmental change.
24 Springs have a significant presence of bryophytes and hence they are ideal habitats for
25 studying their relationship with the environment. We tested whether bryophyte
26 assemblages can be explained with macro-, meso- and micro-ecological variables (i.e.,
27 seasonal climate, altitude, water pH and conductivity) sampling bryophytes from 198
28 semi-natural springs distributed along montane regions in the north-eastern Iberian
29 Peninsula. We tested the influence of environmental variables on bryophyte
30 assemblages in springs using sparse Partial Least Squares (sPLS). Our results show
31 that variability in bryophyte assemblages is explained by seasonal climate (temperature
32 and precipitation from winter, spring, summer and autumn and temperature and
33 precipitation seasonality), altitude and water conductivity. The results obtained by the
34 present study will be useful for predicting bryophyte diversity in springs using simple and
35 easy to obtain variables such as climate, water pH and conductivity.

36 Introduction

37 Bryophytes are very sensitive to environmental conditions (Kapfer *et al.*, 2012), mainly
38 because they do not have a well-developed cuticle or epidermis, and hence most of them
39 can absorb water and nutrients through their entire surface. This is why bryophytes are
40 a very interesting group to study the effects of environmental variability and pollution on
41 community composition. In fact, bryophytes have previously been used as bioindicators
42 of air pollution (Slack, 1990; Suren & Ormerod, 1998; Suren & Duncan, 1999) and water
43 quality (Arts, 1990; Vanderpoorten & Klein 1999; Ceschin *et al.*, 2012).

44 Despite the fact that bryophytes are generally more widely distributed than other plants
45 because of their dispersal capacity (Patiño *et al.*, 2015; Wang *et al.*, 2017), species
46 distribution is very dependent on environmental conditions. The most studied variables
47 include climate (Gignac & Vitt, 1990; Nicholson *et al.*, 1996), water availability (Pentecost
48 & Zhang, 2006), and geology (Pentecost & Zhaohui, 2002; Belland, 2005; Callaghan &
49 Ashton, 2008a) which, through their interaction, determine more local and potentially
50 important environmental variables such as water pH (Tyler *et al.*, 2018) and conductivity
51 in springs and streams where bryophytes can establish (Peñuelas & Sabater, 1987;
52 Vanderpoorten & Palm, 1998; Sabater *et al.*, 2015).

53 One of the habitats that can be dominated by bryophytes is the headwaters of rivers and
54 springs. In these environments, climate, altitude, water pH and conductivity have been
55 suggested to determine which species can become established (Tomaselli *et al.*, 2011;
56 Ceschin *et al.*, 2012; Tessler *et al.*, 2014a; Vieira *et al.*, 2014). However, other micro-
57 and mesoscale environmental variables such as substrate or habitat heterogeneity
58 (Suren, 1996; Strohbach *et al.*, 2009; Vieira *et al.*, 2014), aspect and slope (Pentecost &
59 Zhang, 2006), or disturbance regime of floods and droughts may also be factors
60 explaining variation in bryophyte assemblages in aquatic and semi-aquatic habitats
61 (Suren & Ormerod, 1998; Suren & Duncan, 1999). In similar habitats, such as peatlands,

62 some studies have suggested that bryophyte assemblages depend on gradients of
63 temperature and precipitation (Gignac, 1994), water pH and conductivity (Gignac, 1992)
64 or shade and permafrost (Belland & Vitt, 1995). Other studies carried out in temperate
65 ecosystems of England (Callaghan & Ashton, 2008a), coastal zones of Canada (Belland,
66 2005), and semi-arid zones of Australia (Eldridge & Tozer, 1997) indicate that the
67 different bryophyte assemblages depended on precipitation, soil pH and geology,
68 showing a clear differentiation between communities located over calcareous and non-
69 calcareous lithologies (Callaghan & Ashton, 2008b; Spitale *et al.*, 2009; Virtanen *et al.*,
70 2009). Additionally, bryophyte communities have also been suggested to vary along
71 altitudinal gradients (Wolf, 1993; Andrew & Rodgers, 2003; Bruun *et al.*, 2006; Grytnes
72 *et al.*, 2006; Ah-Peng *et al.*, 2007; Grau *et al.*, 2007; Spitale, 2016), which can be a proxy
73 for differences in topology and radiation amongst others.

74 Notwithstanding all this research on bryophyte ecology, the study of differences in
75 bryophyte assemblages along gradients of altitude, seasonal temperature and
76 precipitation, water pH and conductivity combined, is still very limited in the region of
77 Catalonia and especially in semi-natural or man-made springs (Corbera *et al.*, 2015). In
78 our study region (Catalonia, NE Iberian Peninsula) the ecology of bryophytes has been
79 poorly studied (but see Peñuelas, 1983; Peñuelas & Sabater, 1987), especially
80 concerning the interaction between macro-, meso- and microecological variables.
81 Hence, the aim of this study was to test to what extent macro-, meso- and
82 microecological variables such as climate (seasonal temperatures and precipitation and
83 their seasonality), altitude, and water pH and conductivity can explain variation in
84 bryophyte species assemblage in springs within a relatively large climate and water
85 chemistry gradient. To achieve our aim, we surveyed springs and their bryophyte
86 communities in montane regions of north-eastern Iberian Peninsula (Catalonia, NE
87 Iberian Peninsula). Based on results provided by previous studies, we hypothesised that
88 water pH and conductivity would play a main role determining bryophyte communities of

89 semi-natural springs, due to the direct contact between the bryophytes and the water,
90 and species differences in tolerance ranges for these properties. In addition, because in
91 the Mediterranean climate, summer is the warmest and driest season of the year, we
92 hypothesised that climate from the summer season (dry and warm season) would also
93 significantly explain variability in species composition of spring bryophyte communities.
94 Altitude was also hypothesised to be an important factor for determining spring bryophyte
95 assemblage as it can reflect differences in topography, radiation and climate.

96 **Materials and methods**

97 **Study area and experimental design**

98 We surveyed 198 water springs following a 95 km south-to-north and a 50 km east-to-
99 west gradient along the north-eastern Iberian Peninsula territory (Figure 1), comprising
100 4 different mountain ranges: the Central Littoral mountain range, Montseny-Guilleries,
101 Lluçanès and Garrotxa. The springs were unevenly distributed across the different
102 mountain ranges: 101 springs occurred in the Montseny-Guilleries, 55 in the Central
103 Littoral mountain range, 32 in Lluçanès and 10 in Garrotxa. Climate also differs amongst
104 these mountain ranges. The Central Littoral mountain range has a typical maritime
105 Mediterranean climate with mild summers and winters, while the climate is
106 Mediterranean humid and sub-humid in Montseny-Guilleries and Lluçanès and
107 Mediterranean pre-Pyreneal in Garrotxa (Martín-Vide, 1992), being more continental.
108 Lithologically, the springs of the Central Littoral mountain range are located over different
109 types of granite, except for the eastern side which is dominated by metamorphic rocks
110 (Sabater *et al.*, 2015; Fernández-Martínez *et al.*, 2016). In Montseny, springs are located
111 over granite, metamorphic and calcareous rocks while in Guilleries most of them are over
112 granite. Lluçanès is completely calcareous and Garrotxa is divided into calcareous and
113 volcanic rocks. This rich lithology produces a large variability in the characteristics of

114 water from springs, affecting dissolved ions and thus conductivity and pH (Sabater *et al.*,
115 2015).

116 Most of the springs we surveyed are small human-made constructions to collect water
117 from the underground and release it through a spout (Figure S1). Some of the springs
118 drained water mines (collecting water from aquifers) while others were built in naturally
119 occurring springs. However, in all of the surveyed springs, flowing water did not receive
120 any sanitary treatment. The morphology of our springs did not vary greatly (for a typical
121 example see Figure S1). Almost all of them contained the same microhabitats such as a
122 spout, wet rock walls, soil ground/walls and a sink where the water makes a little pool.
123 Surveys of the springs were carried out during spring and autumn seasons of the years
124 2013 and 2015, and although many other springs were surveyed, only those that had
125 water flowing at the time of the survey were sampled and included in this study. We also
126 excluded from analysis any springs whose water flow was controlled by a tap. Therefore,
127 the springs we sampled contained bryophytes that were continuously receiving water
128 during most of the year, apart from springs in which drought had temporarily interrupted
129 water flow. Water pH and conductivity of the flowing water was analysed in the field with
130 pH-meter (ORION, Thermo-Scientific, Waltham), previously calibrated with standard
131 solutions at pH 7 and 10, while electric conductivity was measured using a conductivity-
132 meter (WTW, Xylem Inc., Weilheim) calibrated at 25°C and with a standard solution of
133 1314 $\mu\text{S cm}^{-1}$.

134 To determine the bryophyte communities (as presence or absence data) we took a
135 sample of the different bryophyte species present in the spring that were either in direct
136 contact with the water flow, receiving water drops or under water (see yellow line in
137 Figure S1). We did not standardise our bryophyte measurements by the total area
138 sampled because, (1) sampled area was lower than 0.9 m² for most of the springs, for
139 which differences in scale were minimal, (2) species-area relationship was missing for a
140 subset of our dataset (Figure S2) and, (3) surveying a predetermined area would have

141 made us miss some species present in the springs. Species number per spring was
142 generally very low, most of them having between 1 and 4 species (Figure S3). We used
143 identification keys in Smith (1978 & 1990) and Casas *et al.*, (2001 & 2004) and follow
144 the nomenclature established by Hill *et al.*, (2006). Some of the samples could not be
145 identified to species level (3%, 18 out of 538 samples). To study the variation in
146 bryophyte assemblages we used those bryophyte species that appeared in more than
147 10 springs (13 species in total), to ensure enough replicates per species.

148 Using the geographical coordinates of the springs, obtained in the field with a GPS, we
149 extracted average seasonal (winter, spring, summer and autumn) temperatures and
150 precipitation, as well as mean annual temperature (MAT) and precipitation (MAP), from
151 the Digital Climatic Atlas of Catalonia (Pons, 1996; Ninyerola *et al.*, 2000; available at
152 <http://www.opengis.uab.cat/acdc/index.htm>) with a resolution of 180 m. Seasonal climate
153 was calculated as the average temperature and total precipitation values of December
154 to February for winter, March to May for spring, June to August for summer and
155 September to November for autumn. Using seasonal climate, we also calculated
156 temperature and precipitation seasonality as the coefficient of variation of seasonal
157 temperatures and precipitation. Altitude was obtained using a GPS in the field.

158 **Statistical analyses**

159 To test whether bryophyte assemblages depend on environmental conditions we
160 performed a sparse Partial Least Squares (sPLS) analysis (Lê Cao *et al.*, 2008), fitted
161 by the regression mode, where water pH and conductivity, seasonal temperatures and
162 precipitation, temperature and precipitation seasonality, and altitude were the
163 environmental matrix of predictors and the binary variables of bryophyte species
164 presence were the response matrix to be predicted. Because the response matrix
165 (species presence) was a binary variable we fitted the model after applying a log-ratio
166 transformation. PLS family methods work by finding the underlying relationships between

167 two matrices (X and Y, predictors and responses) by extracting latent variables to model
168 the covariance structures of the two spaces. PLS models attempt to estimate axes or
169 variates in the X matrix that explains the maximum variance in Y matrix. These methods
170 are particularly useful when the X matrix has more variables than observations, and
171 when high multicollinearity is present among X values. Additionally, sparse PLS methods
172 include a process of selection of the predictor variables to facilitate biological
173 interpretation of the results. The sPLS model was fitted using the “spls” function in
174 *mixOmics* R package (Le Cao *et al.*, 2017). Tuning of the model suggested that
175 extracting more than two variates did not improve the prediction of the response matrix
176 (total Q value saturates at the second variate extracted). To visualise the results of the
177 sPLS analysis, we plotted a Clustered Image Map (CIM), using the “cim” function in
178 *mixOmics*, which shows the correlations between the predictor and the response
179 variables while clustering both based on their similarity on the multivariate space.

180 To look for significant differences of the variables selected by the sPLS analysis among
181 species, we performed analyses of variance (ANOVA) in which we related the
182 environmental variables selected in the sPLS as a function of the species. This analysis
183 allows us to estimate ranges of tolerance for the species and to test whether different
184 species share their ecological niche or not. Differences amongst species were tested
185 using the post-hoc Tukey HSD test. These analyses were done with the 82 springs in
186 which at least one of the most frequent species was found. All ANOVAs accomplished
187 the required statistical assumptions of normality, homoscedasticity and independence of
188 the residuals. All the statistical analysis was performed using R statistical software (R
189 Core Team, 2015).

190 **Results**

191 **Water chemistry and climate of the studied springs**

192 We found a relatively large gradient in water chemistry and climatological features of the
193 surveyed springs. Mean annual temperature (MAT) ranged from 7.3 to 15.7°C, mean
194 annual precipitation (MAP) from 555 to 1180 mm year⁻¹, pH from 5.7 to 8.8, conductivity
195 from 19 to 2160 $\mu\text{S cm}^{-1}$ and altitude from 100 to 1570 m a.s.l. The springs of the Littoral
196 mountain range are located at lower altitudes, being on average, 328 m above sea level
197 (a.s.l.) and therefore, experienced the highest temperatures (13.9°C on average) (Table
198 1). Springs from Lluçanès, Montseny-Guilleries and Garrotxa are found at intermediate
199 altitudes, between 550 and 850 m a.s.l., and recorded temperatures between 11 and
200 12.5°C, although some of Montseny springs experienced relatively high temperatures
201 (up to 15.7°C). Annual precipitation in La Garrotxa is higher than in any other of the
202 studied regions (1049 mm), followed by Montseny-Guilleries, whereas Lluçanès and the
203 Littoral mountain range are drier (Table 1).

204 In general, pH of the springs surveyed was slightly alkaline. However, pH in Garrotxa's
205 springs is higher than those from the other regions studied (Table 1). Conductivity,
206 though, is much more variable than pH. Lluçanès' springs are, on average, those with
207 highest conductivity (1059 $\mu\text{S}\cdot\text{cm}^{-1}$), whereas springs from the Littoral mountain range
208 and Garrotxa are found in the mid-range (around 700 $\mu\text{S}\cdot\text{cm}^{-1}$), and those from
209 Montseny-Guilleries have the lowest conductivity values of all (286 $\mu\text{S}\cdot\text{cm}^{-1}$) (Table 1).

210 **Relationships between species presence and environmental variables**

211 The sPLS model included conductivity, altitude, spring, summer and autumn
212 temperature, winter, spring, summer and autumn precipitation, and temperature and
213 precipitation seasonality as important predictors of bryophyte species assemblages in
214 springs (Figure 2). The first variate explained 51.1% of the variability in the predictors
215 and 12.9% of the variability of the bryophyte assemblages. As shown by Figure S4, no
216 single species or spring was driving the results of our sPLS model. This variate was
217 positively related to seasonal precipitation, altitude, and temperature seasonality and

218 negatively related to conductivity, seasonal temperatures and precipitation seasonality
219 (Figure S5). The second variate explained 28.1% of the variability in the predictors and
220 11.0% of the variability of the bryophyte assemblages. This variate was positively related
221 mainly to summer precipitation and altitude and negatively related mainly to seasonal
222 temperatures and precipitation seasonality.

223 Species were clustered into two big clusters (1 and 2) with two smaller clusters each (a
224 and b, Figure 2). The first cluster of species (1a), formed by *Eucladium verticillatum*,
225 *Pohlia melanodon*, *Didymodon tophaceus* and *Pellia endiviifolia*, was found to be related
226 to warm and especially dry climate and high water conductivity. Their presence was also
227 negatively related to altitude of the springs and temperature seasonality and, less
228 strongly, positively related to precipitation seasonality. *Pallustriella commutata* and
229 *Cratoneuron filicinum*, which formed cluster 1b, were found to be associated with dry and
230 cold climate with low temperature and precipitation seasonality, but with rainy summers
231 and located at high altitude. *Pallustriella commutata* was also positively related to high
232 water conductivity, while no relationship was found with *C. filicinum*.

233 *Kindbergia praelonga*, *Oxyrrhynchium speciosum* and *Conocephalum conicum* formed
234 cluster 2a and were more likely to be found in springs that have a warm and relatively
235 dry climate in spring, and especially in summer, and high water conductivity (except *C.*
236 *conicum*). The species of this cluster were also found in springs of low altitude and low
237 seasonal variability of temperature and precipitation (low continentality). The last cluster
238 of species (2b) was formed by *Plagiomnium undulatum*, *Brachythecium rivulare*, *Bryum*
239 *pseudotriquetrum*, and *Platyhypnidium riparioides*. These species were the group of
240 mosses related to the springs with the coldest and wettest climate, highest temperature
241 seasonality but lowest precipitation seasonality, and highest altitude and having also low
242 water conductivity. *Plagiomnium undulatum*, however, was not so strongly related to cold
243 climate.

244 Of all studied species, *P. endiviifolia* and *C. conicum* were the species with the lower
245 correlations (Figure 2). Table 3 shows the average values per species for the
246 environmental variables included in the sPLS. Significant differences appear amongst
247 species for all of the variables, which further corroborates results from the sPLS. This
248 indicates that ecological niche separation occurs within the environmental variables
249 considered.

250 **Discussion**

251 Our results show that macroclimate, altitude and water conductivity can explain variability
252 in bryophyte assemblages in semi-natural and man-made springs. The distribution of
253 some of the most abundant bryophyte species in our springs is clearly separated into
254 different ecological niches, defined by the environmental variables we considered. In
255 particular, our findings suggest that models predicting species distributions could be
256 improved by including micro- and mesoscale environmental variables such as water
257 conductivity and altitude in addition to macroscale variables such as climate. However,
258 other environmental variables considered in previous studies (Longton *et al.*, 1983;
259 Cattaneo & Fortin, 2000; Tessler *et al.*, 2014; Corbera *et al.*, 2015), such as main ions
260 (Ca^{2+} , Na^+ , Mg^{2+} , K^+ , SO_4^{2-} , Cl^- , PO_4^{3-}), pollutants such as nitrate (Vanderpoorten & Klein,
261 1999), or type of microhabitat and its heterogeneity (Suren, 1996) could help to
262 differentiate the ecological niche of bryophyte species even further. Despite that, the
263 present study clearly shows several general trends in bryophyte assemblages in springs.

264 **Determinants of bryophyte assemblages**

265 The role of precipitation on bryophyte assemblages is evident in previous studies
266 (Eldridge & Tozer, 1997; Gignac, 2001; Callaghan & Ashton, 2008a) although some
267 other papers highlight that the number of rainy days per year is sometimes more
268 important than the total amount of precipitation (Ratcliffe, 1968; Callaghan & Ashton,
269 2008b) because the time that the bryophytes are dry is more determinant than the total

270 amount of water they receive (Pentecost & Zhang, 2006; Proctor *et al.*, 2007). However,
271 this fact may not be so relevant in our study, because most of the sampled springs
272 provide water almost constantly throughout the year. We found that *B. rivulare*, *P.*
273 *riparioides*, *B. pseudotriquetrum*, which are species associated with streams and
274 believed not to tolerate desiccation, are those with the highest affinity for high
275 precipitation (Figure 2, Table 2). However, other species with supposedly equally high
276 affinity for humidity like *P. endiviifolia*, *P. commutata*, *C. filicinum* and *E. verticillatum*
277 (Table 2), do not show such a strong positive association with precipitation, but even
278 show the opposite (Figure 2).

279 Kapfer *et al.* (2012) suggested that the length of time water resides in the aquifers of
280 zones with lower precipitation is longer than in zones with higher precipitation (). A longer
281 time in contact with the rock causes the water to dissolve more salts and increase its
282 conductivity. This could explain the presence of known calcicolous species, associated
283 with high conductivity water, in springs with low precipitation (Corbera *et al.*, 2015)
284 (Figure 2). Despite the fact that bryophytes are able to tolerate wide ranges of
285 temperature, we found different average seasonal temperatures and temperature
286 seasonality (which relates to continentality) per species. This effect of temperature on
287 bryophyte assemblages may also be mediated through its effect on the water availability
288 of the spring. Almost all of our springs had forested areas nearby and which typically
289 consume large quantities of subsoil water, especially under high temperatures
290 (Fernández-Martínez *et al.*, 2014), thus reducing water runoff. Also, continentality
291 increases temperature extremes throughout the year, which can certainly be an
292 important factor favouring certain species versus others, depending on their temperature
293 tolerance ranges.

294 Some studies have shown a great polarization in the distribution of bryophytes in relation
295 to calcareous and non-calcareous lithology (Callaghan & Ashton, 2008a; Virtanen *et al.*,
296 2009) because of big differences in the cellular exchange capacity of Ca^{2+} (Bates, 1982).

297 In our study, species with affinity for calcareous lithology, like *D. tophaceus* and *E.*
298 *verticillatum* (Table 2), have been found in springs with high water conductivity (on
299 average, 972 and 886 $\mu\text{S}\cdot\text{cm}^{-1}$ respectively) while calcifugous species like *B. rivulare*
300 (Longton *et al.*, 1983) have been found in springs with low water conductivity (292 $\mu\text{S}\cdot\text{cm}^{-1}$).
301 However, in contrast to other studies carried out with bryophyte communities from
302 peat lands and in the headwaters of rivers (Gignac, 1992; Tessler *et al.*, 2014), we did
303 not find any significant relationship between water pH and bryophyte assemblages
304 (Figure 2). This fact may be the result of having analysed only the 13 most frequent
305 bryophyte species that were found in the springs. As these species are more frequent,
306 they are also supposed to be more likely to be generalists and to tolerate wider ranges
307 of water pH. Further work, based on more intensive sampling and focused on less
308 frequent species, will help elucidate how water pH, conductivity and other environmental
309 variables (e.g., climate, topography and ionic composition of water) affect bryophyte
310 assemblages in Mediterranean springs. However, our results indicate that water
311 conductivity could be a very suitable variable to include in species distribution models to
312 improve their predictions on humid or sub-humid bryophytes, given the fact that it is
313 relatively easy to obtain (it could be inferred from lithology) and significantly separates
314 bryophyte species.

315 **Conclusions**

316 Our results clearly demonstrate that bryophyte assemblages in semi-natural springs can
317 be explained using widely available variables such as mean seasonal temperatures and
318 precipitation, temperature and precipitation seasonality (intra-annual variation), altitude,
319 and water conductivity of the springs. Our results also point out the need to include micro-
320 and mesoscale variables, in addition to macroscale variables to help improve the
321 prediction of species distribution models.

322 **Acknowledgements**

323 MFM is funded by a postdoctoral subsidy of the University of Antwerp. We acknowledge
324 the Institució Catalana d'Història Natural (ICHN) and the Institut d'Estudis Catalans (IEC)
325 for funding the project. Also, to all the volunteers of the Delegation of the Serralada Litoral
326 Central, Grup de Naturalistes d'Osona and Lluçanès Viu for helping with the survey of
327 the springs. We also thank all those people who helped us to find the springs.

328

329 Taxonomic Additions and Changes: Nil.

330

331 **References**

- 332 **Ah-Peng, C., Chuah-Petiot, M., Descamps-Julien, B., Bardat, J., Stamenoff, P. &**
333 **Strasberg, D. 2007.** Bryophyte diversity and distribution along an altitudinal gradient
334 on a lava flow in La Réunion. *Diversity and Distributions*, 13: 654–62.
- 335 **Andrew, N.R. & Rodgerson, L. 2003.** Variation in invertebrate – bryophyte community
336 structure at different spatial scales along altitudinal gradients. *Journal of Biogeography*,
337 30: 731–46.
- 338 **Arts, G.H.P. 1990.** Aquatic bryophyta as indicators of water quality in shallow pools
339 and lakes in The Netherlands. *Annales Botanici Fennici*, 27: 19–32.
- 340 **Atherton, I.D.M., Bosanquet, S.D.S. & Lawley, M.2010.** *Mosses and liverworts of*
341 *Britain and Ireland: a field guide*. Cardiff: British Bryological Society.
- 342 **Bates, J.W. 1982.** The role of exchangeable calcium in saxicolous calcicole and
343 calcifuge mosses. *New Phytologist*, 90: 239–52.
- 344 **Belland, R.J. 2005.** A multivariate study of moss distributions in relation to
345 environment in the Gulf of St. Lawrence region, Canada. *Canadian Journal of Botany*,
346 83: 243–63.

347 **Belland, R.J. & Vitt, D.H. 1995.** Bryophyte vegetation patterns along environmental
348 gradients in continental bogs. *Ecoscience*, 2: 395–407.

349 **Bruun, H.H., Moen, J., Virtanen, R., Grytnes, J.-A., Oksanen, L. & Angerbjörn, A.**
350 **2006.** Effects of altitude and topography on species richness of vascular plants ,
351 bryophytes and lichens in alpine communities. *Journal of Vegetation Science*, 17: 37–
352 46.

353 **Callaghan, D.A. & Ashton, P.A. 2008a.** Bryophyte distribution and environment
354 across an oceanic temperate landscape. *Journal of Bryology*, 30(1): 23–35.

355 **Callaghan, D.A. & Ashton, P.A. 2008b.** Knowledge gaps in bryophyte distribution and
356 prediction of species-richness. *Journal of Bryology*, 30: 147–58.

357 **Le Cao, K.-A., Rohart, F., Gonzalez, I., Dejean, S., Gautier, B., Bartolo, F., Monget,**
358 **P., Coquery, J., Yao, F-Z. & Liquet, B. 2017.** mixOmics: Omics Data Integration
359 Project. Available at: R package version 6.1.1. [https://CRAN.R-](https://CRAN.R-project.org/package=mixOmics)
360 [project.org/package=mixOmics](https://CRAN.R-project.org/package=mixOmics)

361 **Casas, C. 1958.** Aportaciones a la flora briológica de Cataluña. Musgos y hepáticas
362 del Montseny. *Anales del Instituto Botánico Antonio José Cavanilles de Madrid*, 16:
363 121–226.

364 **Casas, C. 1959.** Aportaciones a la flora briológica de Cataluña. Catálogo de las
365 hepáticas y musgos del Montseny. *Anales del Instituto Botánico Antonio José*
366 *Cavanilles de Madrid*, 17: 21–174.

367 **Casas, C., Brugués, M. & Cros, R.M. 2001.** *Flora de Briòfits dels Països Catalans. I.*
368 *Molses*. Barcelona: Institut d'Estudis Catalans.

369 **Casas, C., Brugués, M. & Cros, R.M. 2004.** *Flora de Briòfits dels Països Catalans. II.*
370 *Hepàtiques i antocerotes*. Barcelona: Institut d'Estudis Catalans.

371 **Cattaneo, A. & Fortin, L. 2000.** Moss distribution in streams of the Quebec Laurentian
372 Mountains. *Canadian Journal Of Botany-Revue Canadienne De Botanique*, 78: 748–
373 752.

374 **Ceschin, S., Aleffi, M., Bisceglie, S., Savo, V. & Zuccarello, V. 2012.** Aquatic

375 bryophytes as ecological indicators of the water quality status in the Tiber River basin
376 (Italy). *Ecological Indicators*, 14: 74–81.

377 **Corbera, J., Fernández-Martínez, M., Jover, M., Torner, G., Calpe, M., Ciurana, O.**
378 **& Sabater, F. 2015.** Els briòfits de les fonts de la Serralada Litoral Central: composició
379 específica i efecte dels paràmetres ambientals en la seva distribució. *L'Atzavara*, 25:
380 105–16.

381 **Eldridge, D.J. & Tozer, M.E. 1997.** Environmental factors relating to the distribution of
382 terricolous bryophytes and lichens in semi-arid eastern Australia. *Bryologist*, 100: 28–
383 39.

384 **Fernández-Martínez, M., Bagaria, G., Raya, I., Preece, C., Margalef, O. & Domene,**
385 **X. 2016.** Terra Negra: un sòl sobre pissarres al Montnegre. *L'Atzavara*, 26: 67–72.

386 **Fernández-Martínez, M., Vicca, S., Janssens, I.A., Luysaert, S., Campioli, M.,**
387 **Sardans, J., Estiarte, M. & Peñuelas, J. 2014.** Spatial variability and controls over
388 biomass stocks, carbon fluxes and resource-use efficiencies in forest ecosystems.
389 *Trees, Structure and Function*, 28: 597–611.

390 **Gignac, L.D. 1992.** Niche structure, resource partitioning , and species interactions of
391 mire bryophytes relative to climatic and ecological gradients in western Canada.
392 *American Bryological and Lichenological Society*, 95: 406–18.

393 **Gignac, L.D. 1994.** Peatland species preferences: An overview of our current
394 knowledge base. *Wetlands*, 14: 216–22.

395 **Gignac, D. 2001.** New frontiers in bryology and lichenology: bryophytes as indicators
396 of climate change. *Bryologist*, 104: 410–20.

397 **Gignac, L.D. & Vitt, D. 1990.** Habitat limitations of Sphagnum along climatic, chemical
398 and physical gradients in mires of western Canada. *Bryologist*, 93: 7–22.

399 **Grau, O., Grytnes, J.A. & Birks, H.J.B. 2007.** A comparison of altitudinal species
400 richness patterns of bryophytes with other plant groups in Nepal, Central Himalaya.
401 *Journal of Biogeography*, 34: 1907–15.

402 **Grytnes, J.A., Heegaard, E. & Ihlen, P.G. 2006.** Species richness of vascular plants,

403 bryophytes, and lichens along an altitudinal gradient in western Norway. *Acta*
404 *Oecologica*, 29: 241–46.

405 **Hill, M.O., Bell, N., Bruggeman-Nannenga, M., Brugués, M., Cano, M.J., Enroth, J.,**
406 **Flatberg, K.I., Frahm, J.-P., Gallego, M.T., Garilleti, R., Guerra, J., Hedenäs, L.,**
407 **Holyoak, D.T., Ignatov, M.S., Lara, F., Mazimpaka, V., Muñoz, J., & Garilleti, R.**
408 **2006.** An annotated checklist of the mosses of Europe and Macaronesia. *Journal of*
409 *Bryology*, 28: 198–267.

410 **Kapfer, J., Audorff, V., Beierkuhnlein, C. & Hertel, E. 2012.** Do bryophytes show a
411 stronger response than vascular plants to interannual changes in spring water quality?
412 *Freshwater Science*, 31: 625–35.

413 **Lê Cao, K.-A., Rossouw, D., Robert-Granié, C. & Besse, P. 2008.** A Sparse PLS for
414 Variable Selection when Integrating Omics Data : Statistical Applications in Genetics
415 and Molecular Biology. *Statistical Applications in Genetics and Molecular Biology*, 7: 1–
416 35.

417 **Longton, R.E., Schuster, R.M., Glime, J.M. & Vitt, D.H. 1983.** Reproductive Biology:
418 a comparison of bryophyte species diversity and niche structure of montane streams
419 and stream banks. *Canadian Journal of Botany*, 1: 386–462.

420 **Martín-Vide, J. 1992.** El clima. In: C. Carreras, ed. *Geografia General dels Països*
421 *Catalans. Enciclopèdia Catalana*. Barcelona, pp. 1–110.

422 **Nicholson, B.J., Gignac, L.D. & Bayley, S.E. 1996.** Peatland distribution along a
423 north-south transect in the Mackenzie River Basin in relation to climatic and
424 environmental gradients. *Vegetatio*, 126: 119–33.

425 **Ninyerola, M., Pons, X., Roure, J.M., Ninyerola, M., Pons, X. & Roure, J.M. 2000.** A
426 methodological approach of climatological modelling of air temperature and
427 precipitation through GIS techniques, A methodological approach of climatological
428 modelling of air temperature and precipitation through GIS techniques. *International*
429 *Journal of Climatology*, 20: 1823, 1823–41,.

430 **Patiño, J., Sólymos, P., Carine, M., Weigelt, P., Kreft, H. & Vanderpoorten, A.**

431 **2015.** Island floras are not necessarily more species poor than continental ones.
432 *Journal of Biogeography*, 42: 8–10.

433 **Pentecost, A. & Zhang, Z. 2006.** Response of bryophytes to exposure and water
434 availability on some European travertines. *Journal of Bryology*, 28(1): 21–6.

435 **Pentecost, A. & Zhaohui, Z. 2002.** Bryophytes from some travertine-depositing sites
436 in France and the U.K.: relationships with climate and water chemistry. *Journal of*
437 *Bryology*, 24: 233–41.

438 **Peñuelas, J. 1983.** La qualitat de les aigües del riu Muga: Dades físiques, químiques i
439 biològiques. *Annals de l'Institut d'Estudis Empordanesos*, 16: 33–49.

440 **Peñuelas, J. & Sabater, F. 1987.** Distribution of macrophytes in relation to
441 environmental factors in the Ter River, NE Spain. *Internationale Revue der gesamten*
442 *Hydrobiologie und Hydrographie*, 72: 41–58.

443 **Pons, X. 1996.** Estimación de la radiación solar a partir de modelos digitales de
444 elevaciones. Propuesta metodológica. In: *Modelos y sistemas de información*
445 *geográfica*, pp. 87–97.

446 **Proctor, M.C.F., Oliver, M.J., Wood, A.J., Alpert, P., Stark, L.R., Cleavitt, N.L.**
447 **& Mishler, B.D. 2007.** Desiccation-tolerance in bryophytes: a review. *Bryologist*, 110:
448 595–621.

449 **R Core Team. 2015.** R: A Language and Environment for Statistical Computing,
450 version 3.2.3. Available at: <https://cran.r-project.org>.

451 **Ratcliffe, D.A. 1968.** An ecological account of Atlantic Bryophytes in the British Isles.
452 *New Phytologist*, 67: 365–439.

453 **Sabater, F., Fernández-Martínez, M., Corbera, J., Calpe, M., Torner, G., Cano, O.,**
454 **Corbera, G., Ciurana, O. & Parera, J.M. 2015.** Caracterització hidrogeoquímica de les
455 fonts de la Serralada Litoral Central en relació a la litologia i als factors ambientals.
456 *L'Atzavara*, 25: 93–104.

457 **Slack, N.G. 1990.** Bryophytes and ecological niche theory. *Botanical Journal of the*
458 *Linnean Society*, 104: 187–213.

459 **Smith, A.J.E. 1978.** *The moss flora of Britain and Ireland*. 1st edn. Cambridge:
460 Cambridge University Press.

461 **Smith, A.J.E. 1990.** *The liverworts of Britain and Ireland*. Cambridge: Cambridge
462 University Press.

463 **Spitale, D. 2016.** The interaction between elevational gradient and substratum reveals
464 how bryophytes respond to the climate. *Journal of Vegetation Science*, 27: 844–53.

465 **Spitale, D., Petraglia, A. & Tomaselli, M. 2009.** Structural equation modelling detects
466 unexpected differences between bryophyte and vascular plant richness along multiple
467 environmental gradients. *Journal of Biogeography*, 36: 745–55.

468 **Strohbach, M., Audorff, V. & Beierkuhnlein, C. 2009.** Drivers of plant species
469 composition in siliceous spring ecosystems: groundwater chemistry, catchment traits or
470 spatial factors? *Journal of Limnology*, 68: 375–84.

471 **Suren, A.M. 1996.** Bryophyte distribution patterns in relation to macro-, meso-, and
472 micro-scale variables in South Island, New Zealand streams. *New Zealand Journal of*
473 *Marine and Freshwater Research*, 30: 501–23.

474 **Suren, A.M. & Duncan, M.J. 1999.** Rolling stones and mosses: effect of substrate
475 stability on bryophyte communities in streams. *Journal of the North American*
476 *Benthological Society*, 18: 457–67.

477 **Suren, A.M. & Ormerod, S.J. 1998.** Aquatic bryophytes in Himalayan streams: Testing
478 a distribution model in a highly heterogeneous environment. *Freshwater Biology*, 40:
479 697–716.

480 **Tessler, M., Truhn, K.M., Bliss-Moreau, M. & Wehr, J.D. 2014.** Diversity and
481 distribution of stream bryophytes : does pH matter? *Freshwater Science*, 33: 778–87.

482 **Tomaselli, M., Spitale, D. & Petraglia, A. 2011.** Phytosociological and ecological
483 study of springs in Trentino (south-eastern Alps, Italy). *Journal of Limnology*, 70: 25–
484 53.

485 **Tyler, T., Bengtsson, F., Dahlberg, C.J., Lönnell, N., Hallingbäck, T. & Reitalu, T.**
486 **2018.** Determinants of bryophyte species composition and diversity on the Great Alvar

487 of Öland , Sweden. *Journal of Bryology*, 40(1): in press.

488 **Vanderpoorten, A. & Klein, J.P. 1999.** Variations of aquatic bryophyte assemblages
489 in the Rhine Rift related to water quality. 2. The waterfalls of the Vosges and the Black
490 Forest. *Journal of Bryology*, 21: 109–15.

491 **Vanderpoorten, A. & Palm, R. 1998.** Canonical variables of aquatic bryophyte
492 combinations for predicting water trophic level. *Hydrobiologia*, 386: 85–93.

493 **Vieira, C., Aguiar, F.C. & Ferreira, M.T. 2014.** The relevance of bryophytes in the
494 macrophyte-based reference conditions in Portuguese rivers. *Hydrobiologia*, 737: 245–
495 64.

496 **Virtanen, R., Ilmonen, J., Paasivirta, L. & Muotka, T. 2009.** Community concordance
497 between bryophyte and insect assemblages in boreal springs: a broad-scale study in
498 isolated habitats. *Freshwater Biology*, 54: 1651–62.

499 **Wang, J., Vanderpoorten, A., Hagborg, A., Goffinet, B., Laenen, B. & Patiño, J.**
500 **2017.** Evidence for a latitudinal diversity gradient in liverworts and hornworts. *Journal of*
501 *Biogeography*, 44: 487–8.

502 **Wolf, J.H.D. 1993.** Diversity Patterns and Biomass of Epiphytic Bryophytes and
503 Lichens Along an Altitudinal Gradient in the Northern Andes. *Annals of the Missouri*
504 *Botanical Garden*, 80: 928–60.

505

506 **Figure captions**

507

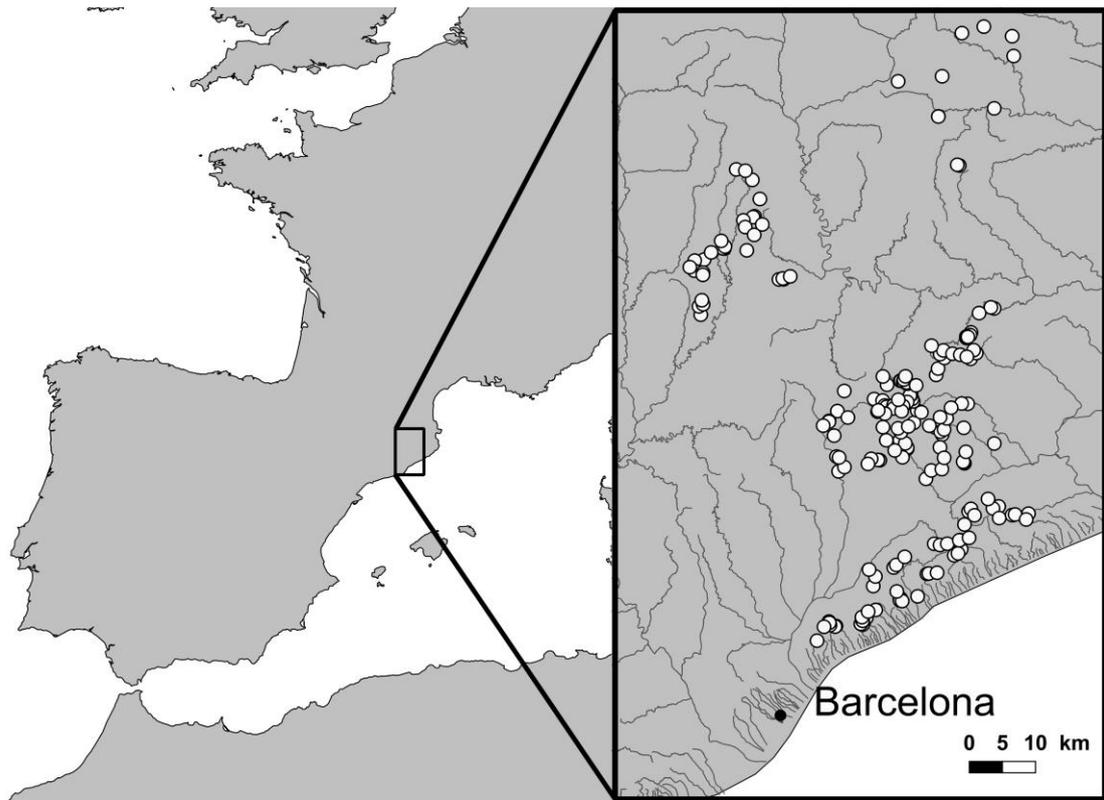
508 **Figure 1.** Map showing the location of the surveyed springs (white dots).

509

510 **Figure 2.** Clustered image map of the sPLS model showing the correlations between
511 environmental predictors and presence of bryophyte species. Species and predictors
512 have been clustered according to their similarity forming groups that we split using black
513 lines. Abbreviations: T indicates temperature, P indicates precipitation, and seasons
514 were indicated in lowercase letters as winter (w), spring (sp), summer (sm) and autumn
515 (a). PS and TS indicate seasonality of precipitation and temperature respectively.

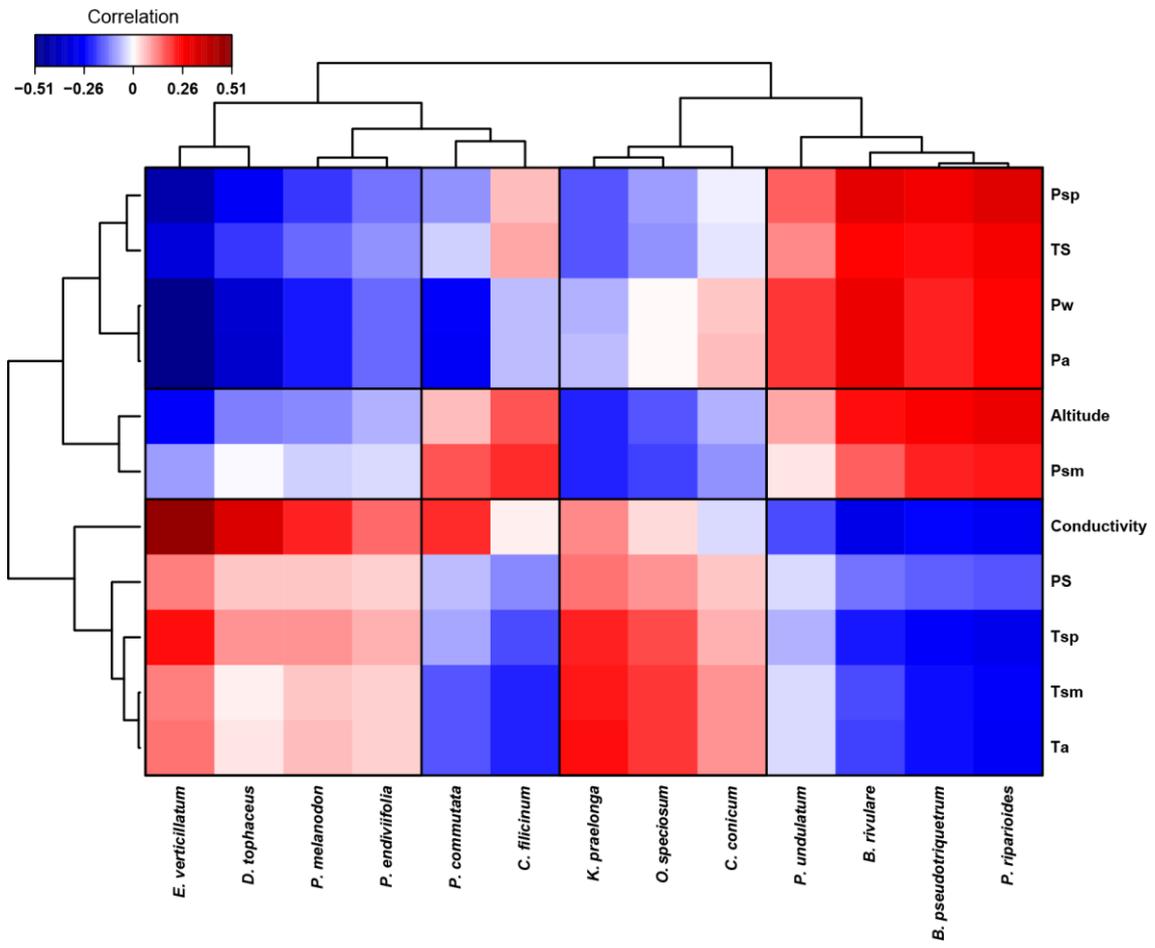
516

517 **Figure 1.**



518

519



522 **Table 1.** Average values for altitude, climate (mean \pm standard error of the mean [SE]), water pH
 523 and conductivity and bryophyte species richness of the springs and average values per region.
 524 Lluçanès is a calcareous region, Montseny and the Central Littoral mountain range have granitic
 525 and metamorphic lithology and Garrotxa springs are located over volcanic and calcareous rocks.
 526 Acronyms are MAT for mean annual temperature and MAP for mean annual precipitation. Units
 527 are degrees Celsius for MAT, mm \cdot year⁻¹ for MAP, μ S \cdot cm⁻¹ for conductivity and metres for altitude.

	Total	Lluçanès	Montseny - Guilleries	Central Littoral	Garrotxa
Springs	198	32	101	55	10
MAT	12.01 \pm 0.13	11.38 \pm 0.12	11.18 \pm 0.30	13.90 \pm 0.16	12.11 \pm 0.52
MAP	863 \pm 11	766 \pm 22	947 \pm 19	728 \pm 14	1049 \pm 31
Altitude	660 \pm 24	703 \pm 29	828 \pm 53	328 \pm 25	640 \pm 109
pH	7.17 \pm 0.04	7.10 \pm 0.09	7.18 \pm 0.11	7.09 \pm 0.10	7.65 \pm 0.13
Conductivity	546 \pm 32	1059 \pm 100	286 \pm 50	701 \pm 74	688 \pm 186
Sp. richness	2.71 \pm 0.12	3.16 \pm 0.44	2.75 \pm 0.23	2.05 \pm 0.21	4.27 \pm 1.31

528
 529

535 **Table 3.** Species average (\pm standard error of the mean) values for environmental variables that
536 were selected in the sPLS for explaining variability in species distribution. Letters (a, b, c, d, e, f)
537 indicate groups according to the post-hoc Tukey HSD test for multiple comparisons. All ANOVA
538 tables were statistically significant at the <0.001 level. Units were metres for altitude, mm year^{-1}
539 for precipitation variables and Celsius degrees for temperature variables and ($\mu\text{S}\cdot\text{cm}^{-1}$) for
540 conductivity. Precipitation seasonality (PS) and temperature seasonality (TS) were unit less.
541 Abbreviations: T indicates temperature, P indicates precipitation, and seasons were indicated in
542 lowercase letters as winter (w), spring (sp), summer (sm) and autumn (a).

Species	Tsp		Tsm		Ta	
<i>Bryum pseudotriquetrum</i>	6.98 \pm 0.58	a	16.1 \pm 0.6	a	10.2 \pm 0.5	a
<i>Brachythecium rivulare</i>	7.50 \pm 0.45	a	16.6 \pm 0.5	a	10.6 \pm 0.4	ab
<i>Platyhypnidium riparioides</i>	7.72 \pm 0.41	ab	16.7 \pm 0.4	ab	10.7 \pm 0.4	abc
<i>Plagiomnium undulatum</i>	9.45 \pm 0.32	bc	18.5 \pm 0.3	bc	12.2 \pm 0.3	cd
<i>Cratoneuron filicinum</i>	9.58 \pm 0.30	cd	18.8 \pm 0.3	c	12.3 \pm 0.3	d
<i>Oxyrrhynchium speciosum</i>	10.15 \pm 0.25	cdc	19.2 \pm 0.3	cd	13.0 \pm 0.2	def
<i>Palustriella commutata</i>	9.88 \pm 0.46	cdc	19.4 \pm 0.5	cd	12.5 \pm 0.4	bcde
<i>Pellia endiviifolia</i>	10.45 \pm 0.27	cdc	19.6 \pm 0.3	cd	13.2 \pm 0.2	def
<i>Conocephalum conicum</i>	11.03 \pm 0.56	cdc	20.0 \pm 0.6	cd	13.6 \pm 0.50	def
<i>Pohlia melanodon</i>	10.88 \pm 0.58	cdc	20.1 \pm 0.6	cd	13.6 \pm 0.5	def
<i>Eucladium verticillatum</i>	11.06 \pm 0.28	c	20.2 \pm 0.3	d	13.7 \pm 0.3	ef
<i>Didymodon tophaceus</i>	10.91 \pm 0.45	cdc	20.3 \pm 0.5	cd	13.5 \pm 0.4	def
<i>Kindbergia praelonga</i>	11.84 \pm 0.61	dc	20.5 \pm 0.6	cd	14.8 \pm 0.6	f

	Pw		Psm		Pa	
<i>Kindbergia praelonga</i>	159.2 \pm 11.8	abc	150.9 \pm 15.7	a	252.5 \pm 10.7	abcd
<i>Oxyrrhynchium speciosum</i>	182.1 \pm 4.9	bcd	175.0 \pm 6.5	ab	263.1 \pm 4.4	bc
<i>Pohlia melanodon</i>	150.8 \pm 11.3	ab	183.2 \pm 15.0	abc	233.6 \pm 10.2	ab
<i>Conocephalum conicum</i>	196.5 \pm 10.9	bcde	188.5 \pm 14.4	abcd	267.8 \pm 9.8	abcde
<i>Plagiomnium undulatum</i>	201.4 \pm 6.3	cde	194.3 \pm 8.3	abc	278.0 \pm 5.7	cde
<i>Brachythecium rivulare</i>	220.0 \pm 8.8	e	196.7 \pm 11.6	abcd	293.7 \pm 7.9	de
<i>Eucladium verticillatum</i>	146.3 \pm 5.5	a	199.2 \pm 7.4	abc	234.9 \pm 5.0	a
<i>Pellia endiviifolia</i>	177.5 \pm 5.2	bc	200.6 \pm 6.9	abc	260.4 \pm 4.7	bc
<i>Didymodon tophaceus</i>	138.6 \pm 8.8	a	210.0 \pm 11.6	abcd	226.6 \pm 7.9	a
<i>Platyhypnidium riparioides</i>	219.2 \pm 8.0	e	216.4 \pm 10.6	bcd	295.6 \pm 7.2	e
<i>Bryum pseudotriquetrum</i>	221.1 \pm 11.3	de	218.9 \pm 15.0	abcd	303.3 \pm 10.2	e
<i>Cratoneuron filicinum</i>	191.3 \pm 5.9	bcde	221.7 \pm 7.8	cd	269.5 \pm 5.3	bcde
<i>Palustriella commutata</i>	155.7 \pm 9.0	ab	248.0 \pm 11.9	d	244.7 \pm 8.1	ab

543

544

545 **Table 3.** Continuation.

Species	Psp	PS	TS
<i>Kindbergia praelonga</i>	18.10 ± 1.01 ^a	0.25 ± 0.012 ^{cd}	0.40 ± 0.03 ^a
<i>Pohlia melanodon</i>	19.88 ± 0.96 ^{abc}	0.24 ± 0.011 ^{abcd}	0.48 ± 0.02 ^{abc}
<i>Eucladium verticillatum</i>	20.03 ± 0.47 ^{ab}	0.25 ± 0.006 ^d	0.48 ± 0.01 ^{ab}
<i>Didymodon tophaceus</i>	20.03 ± 0.75 ^{ab}	0.25 ± 0.009 ^d	0.50 ± 0.02 ^{abc}
<i>Oxyrrhynchium speciosum</i>	21.24 ± 0.42 ^{abc}	0.23 ± 0.005 ^{bcd}	0.50 ± 0.01 ^{bc}
<i>Pellia endiviifolia</i>	22.06 ± 0.44 ^{bcd}	0.22 ± 0.005 ^{abcd}	0.50 ± 0.01 ^{bc}
<i>Conocephalum conicum</i>	22.45 ± 0.93 ^{abcd}	0.21 ± 0.011 ^{abcd}	0.48 ± 0.02 ^{abc}
<i>Palustriella commutata</i>	22.63 ± 0.77 ^{bcd}	0.24 ± 0.009 ^{abcd}	0.54 ± 0.02 ^{bcde}
<i>Plagiomnium undulatum</i>	23.41 ± 0.53 ^{cdc}	0.21 ± 0.006 ^{abc}	0.53 ± 0.01 ^{bcd}
<i>Cratoneuron filicinum</i>	23.96 ± 0.50 ^{dc}	0.20 ± 0.006 ^a	0.54 ± 0.01 ^{cd}
<i>Brachythecium rivulare</i>	24.89 ± 0.75 ^{dc}	0.20 ± 0.009 ^{ab}	0.61 ± 0.02 ^{de}
<i>Platyhypnidium riparioides</i>	25.36 ± 0.68 ^c	0.20 ± 0.008 ^{ab}	0.60 ± 0.02 ^{de}
<i>Bryum pseudotriquetrum</i>	25.84 ± 0.96 ^c	0.20 ± 0.011 ^{abc}	0.64 ± 0.02 ^e

546

Species	Conductivity	Altitude
<i>Platyhypnidium riparioides</i>	227.9 ± 82.4 ^a	893.5 ± 54.3 ^{cd}
<i>Brachythecium rivulare</i>	291.9 ± 90.3 ^{ab}	890.9 ± 59.4 ^{cd}
<i>Plagiomnium undulatum</i>	300.0 ± 64.7 ^a	670.8 ± 42.6 ^{abc}
<i>Bryum pseudotriquetrum</i>	300.7 ± 116.6 ^{ab}	1077.6 ± 76.7 ^d
<i>Conocephalum conicum</i>	429.5 ± 112.0 ^{ab}	448.7 ± 3.7 ^a
<i>Oxyrrhynchium speciosum</i>	505.1 ± 50.5 ^{ab}	551.7 ± 33.2 ^a
<i>Cratoneuron filicinum</i>	560.3 ± 60.9 ^{ab}	755.5 ± 40.1 ^{bc}
<i>Kindbergia praelonga</i>	582.3 ± 121.8 ^{abc}	380.5 ± 80.1 ^a
<i>Pellia endiviifolia</i>	611.9 ± 53.5 ^b	589.6 ± 35.2 ^{ab}
<i>Palustriella commutata</i>	717.9 ± 92.7 ^{bc}	669.2 ± 61.0 ^{abc}
<i>Eucladium verticillatum</i>	885.5 ± 57.1 ^c	516.7 ± 37.6 ^a
<i>Didymodon tophaceus</i>	971.9 ± 90.3 ^c	555.2 ± 59.4 ^{ab}
<i>Pohlia melanodon</i>	1122.3 ± 116.6 ^c	558.6 ± 76.7 ^{ab}

547

548 **Supplementary information**

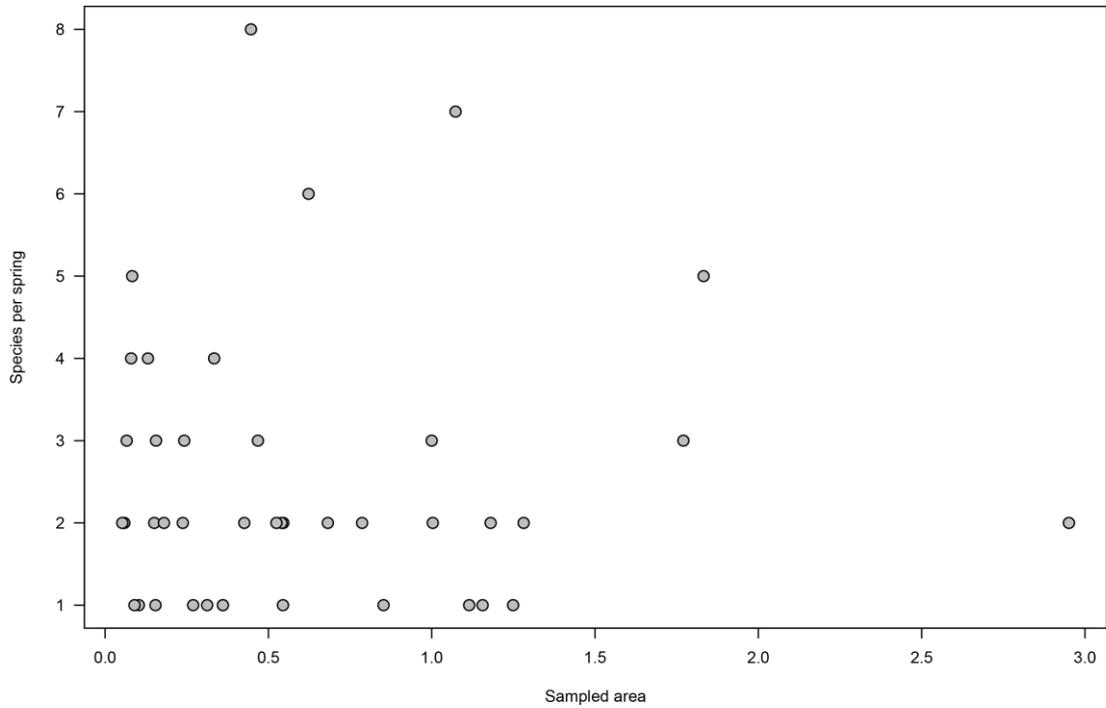
549 **Figure S1:** Picture of a spring surveyed. The yellow line indicates the sampling zone, coinciding
550 with the area under the influence of water.



551

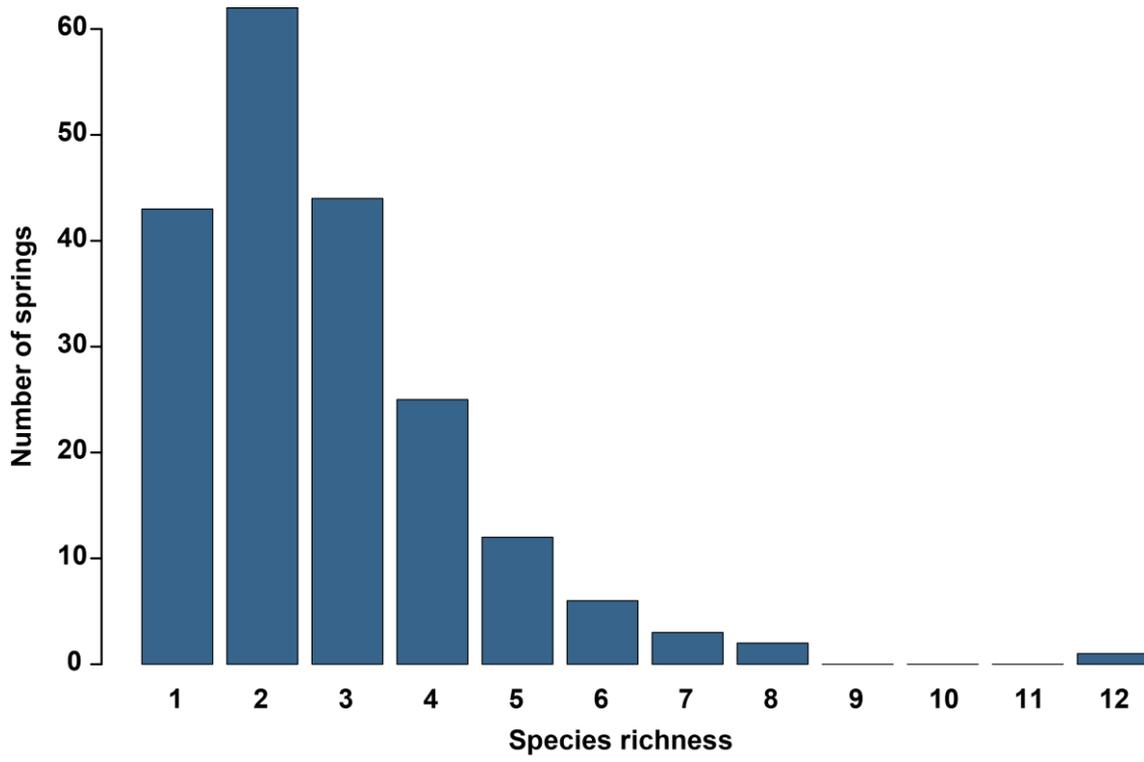
552

553 **Figure S2:** Graph showing species richness as a function of the area sampled in a subset of 41
554 sampled springs. Units of sample area are m⁻². No significant correlation between species number
555 and area was found.



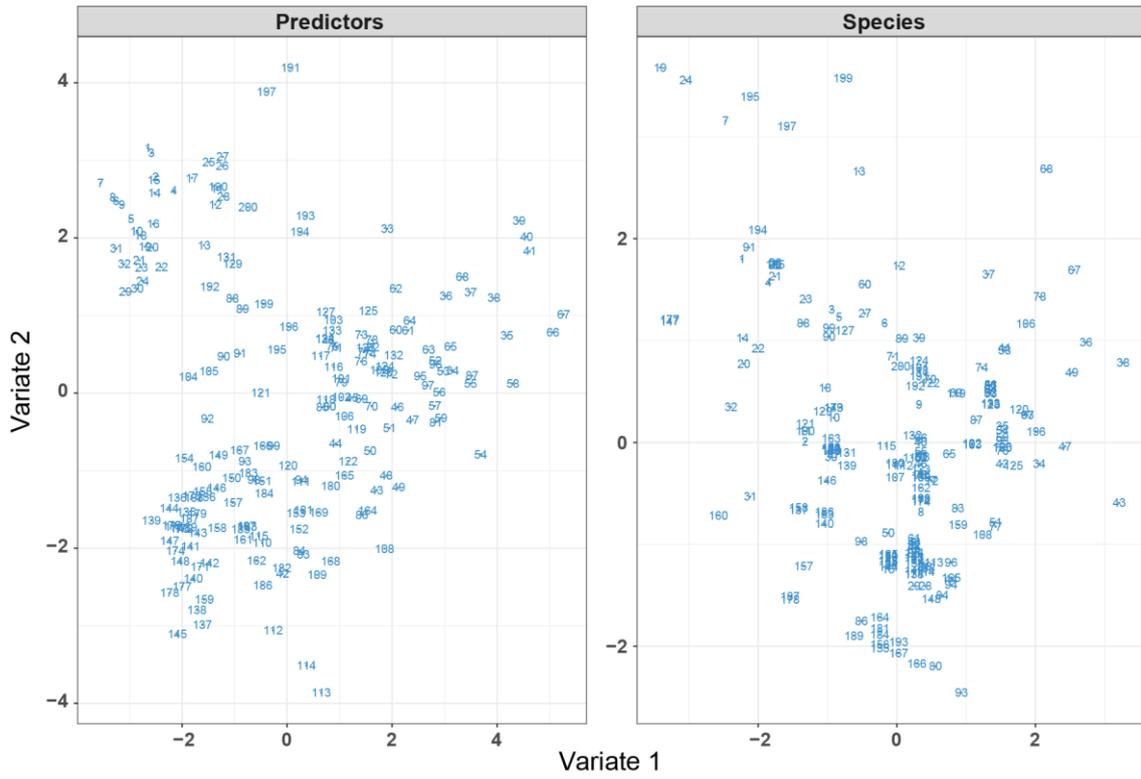
556

557 **Figure S3.** Histogram showing species richness of the studied springs.



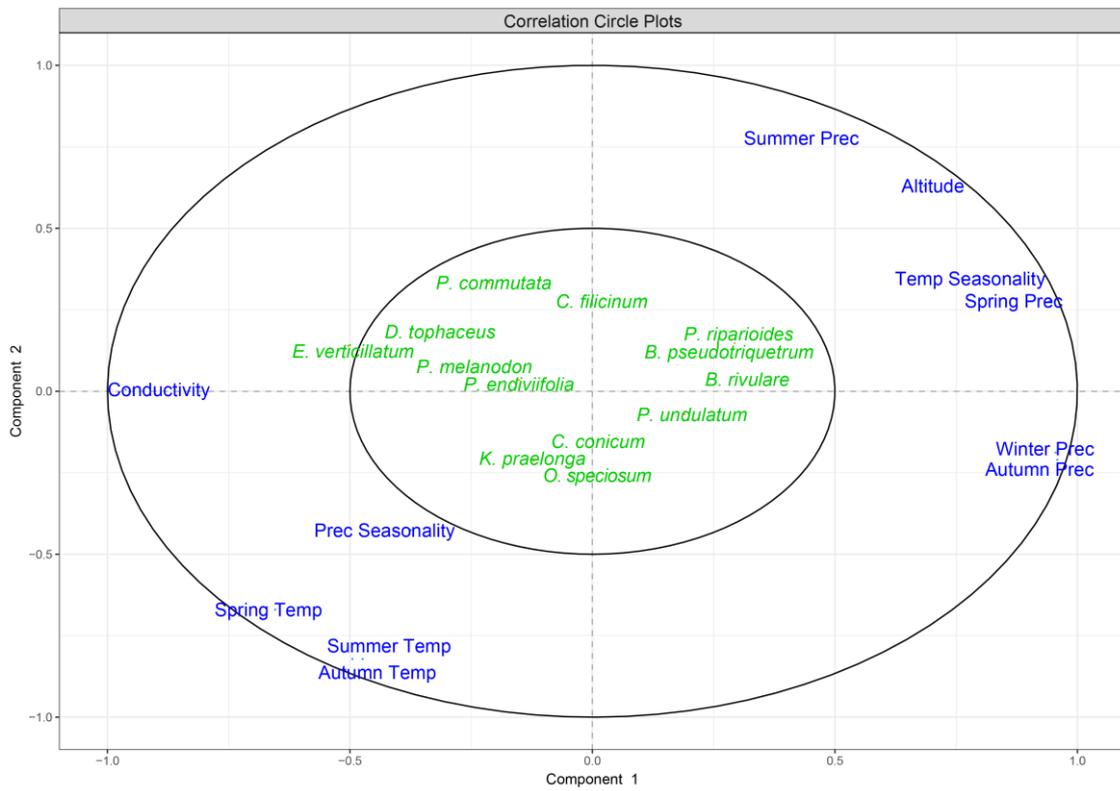
558

559 **Figure S4:** Site scores of the sPLS relating environmental variables (predictors) and bryophyte
560 presence (species). None of the variates were driven by single sites or species.



561

562 **Figure S5:** Correlation circle plot showing the results of the sPLS model relating environmental
563 variables (predictors – in blue) and bryophyte presence (species – in green). Prec: precipitation,
564 Temp: temperature.



565

566