

1 **Shrub encroachment leads to accumulation of C, N, and P in grassland soils and**
2 **alters C:N:P stoichiometry – a meta-analysis**

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31 **Abstract**

32 Soil C:N:P stoichiometry is a sensitive indicator for nutrient balance. Shrub
33 encroachment into grasslands could change the concentrations and stoichiometry of soil
34 carbon (C), nitrogen (N), and phosphorus (P), but the general patterns of these changes
35 remain unclear. With a meta-analysis of a global dataset covering 344 observations
36 from 68 studies, we examined how grassland soil C:N:P stoichiometry responds to
37 shrub encroachment under various environmental conditions. The magnitude of the
38 impact of shrub encroachment on soil organic carbon (SOC), total nitrogen (TN), and
39 total phosphorus (TP) concentrations and their stoichiometry varied with climate, soil
40 texture type, and soil layer. Overall, shrub encroachment significantly increased the
41 concentrations of soil C (+29%), N (+25%), and P (+20%), and it significantly
42 increased soil C:N (+5%), C:P (+12%), and N:P (+6%). Increases in SOC and TN
43 concentrations mainly occurred in dry sub-humid and semi-arid zones. Regarding
44 different soil layers, the increases in SOC and TN concentrations and in the C:N, C:P,
45 and N:P ratios with shrub encroachment were greater in the topsoil (0-20 cm depth)
46 than in deeper soil layers. The concentrations of SOC, TN, and TP all increased
47 significantly with shrub encroachment in the soil texture types, such as clay, sandy loam,
48 and sand. Shrub encroachment increased C:N, C:P, and N:P in acidic and neutral soils.
49 The magnitude of the change in soil C:N was negatively correlated with the duration of
50 shrub encroachment, an effect linked to a greater increase in soil TN than in SOC
51 concentrations with longer durations of encroachment. Our global meta-analysis
52 indicates that shrub encroachment into grasslands increases soil C sequestration and
53 soil C:N, C:P, and N:P. Soil stoichiometric shifts in shrub-encroached grasslands are
54 relatively sensitive to environmental factors, such as soil texture, soil pH, and climate.
55 These findings help us to better understand the effects of shrub encroachment on
56 biogeochemical cycling, functioning, and services in grasslands across a broad range of
57 spatio-temporal scales.

58 **Keywords** global data analysis; shrub encroachment; soil stoichiometry; soil carbon;
59 soil nitrogen; soil phosphorus

60 **Introduction**

61 Carbon (C) stocks in soils far exceed global carbon stocks in vegetation and the
62 atmosphere combined (Lehmann and Kleber, 2015), with grasslands accounting for
63 about one-third of all terrestrial ecosystem carbon stocks (Zhao et al., 2020). Therefore,
64 grassland soils play an important role in the global C cycle (Guo and Gifford, 2002).
65 Over the past century, global grasslands have been invaded by woody plants, especially
66 shrubs, as a result of global changes such as climate warming (Ehleringer, 2005),
67 changes in precipitation (Criado et al., 2020; Berry and Kulmatiski, 2017), elevated
68 CO₂ concentration (Buitenwerf et al., 2012), and human activities including intentional
69 fires (Smit et al., 2016) and overgrazing (Tjelele, 2014). The shifts from grasses to
70 woody plants involve changes in vegetation cover (Nepstad et al., 1994; Jackson et al.,
71 2002) and in soil C sequestration (Briggs et al., 2005). Shrub encroachment may
72 increase C and nitrogen (N) pools in grasslands (Li et al., 2018), but some studies
73 indicate a reduction in soil organic carbon and nitrogen content with woody plant
74 encroachment (Jackson et al., 2002). Therefore, the role of woody plant encroachment
75 in driving grassland C and N cycling remains largely uncertain.

76 Woody plants can concentrate organic matter beneath their canopies and change
77 the microbial biomass, root biomass, litter production, soil characteristics, and
78 microclimate (Binkley et al., 1998; Schlesinger and Pilmanis, 1998), with consequences
79 for soil C sequestration (Connin et al., 1997) and concentration of soil total phosphorus
80 (Mogashoa et al., 2021). Concentrations of soil organic C (SOC) tend to increase after
81 shrub encroachment because of increases in net primary production (NPP; Boutton et
82 al., 2009) and a lower decomposition rate for shrub litter than for grasses (Montané et
83 al., 2010). Furthermore, N-fixing shrubs can lead to a large accumulation of available
84 N (Buhlmann et al., 2014) and increase soil C storage (Connin et al., 1997). Ward et al.
85 (2018) found that soil nutrient concentrations are related to shrub size rather than shrub
86 species, and that large shrubs can increase soil nutrient concentrations in a semi-arid
87 savanna. Soil organic C (SOC) and total nitrogen (TN) concentrations have been
88 reported to increase with increasing shrub island size (Connell et al., 2021) but soil

89 organic C concentration declined with increasing shrub encroachment age (Brantley et
90 al., 2010).

91 Moreover, the impacts of shrub encroachment on soil total carbon (TC) and total
92 nitrogen (TN), and soil total phosphorus (TP) concentrations can be strongly dependent
93 on the climate and soil conditions. For example, edaphic properties significantly affect
94 the size, distribution density, and pattern of shrubs in arid and semi-arid systems
95 (Hughes et al., 2006). In addition, soil nutrient concentrations are related to soil texture,
96 which can provide physical protection of organic matter inputs (Boutton et al., 2009).
97 Moreover, Jackson et al. (2002) compared the C and N budgets of six pairs of shrub-
98 encroached and unencroached grasslands along an annual precipitation gradient of 200
99 to 1100 mm and found a negative correlation between precipitation and changes in soil
100 C and N under shrub-encroached conditions. However, here are still many different
101 opinions on which soil texture and climate is conducive to the sequestration of soil
102 nutrients.

103 Ecological stoichiometry (i.e. element ratios) provides a framework for
104 understanding the relationship between individual elements and how elements
105 are coupled by exploring nutrients cycling in the form of element ratios (Sterner and
106 Elser, 2002). Urbina et al. (2020) found that shrub encroachment slows biogeochemical
107 cycling through the change in plant and soil stoichiometry. C:N:P stoichiometry could
108 be used as a major indicator for soil nutrient limitation and availability, helping to
109 explain crucial ecological processes (Cleveland and Liptzin, 2007; Reich and Oleksyn,
110 2004). For example, the soil C:N ratio could reflect how the decomposition rate of
111 organic matter responds to shrub encroachment, thus enabling predictions of changes
112 in the soil C and N cycles (Wang and Yu, 2008). In addition, ecosystem N and P status
113 and availability may mediate ecosystem responses to climate change (Wieder et al.,
114 2015; Terrer et al., 2019). Previous studies have shown that soil C, N, and P
115 concentrations can respond to woody plant encroachment, but few studies have
116 involved quantifying the responses of C:N:P stoichiometry to shrub encroachment. The
117 impacts of shrub encroachment on soil C:N:P stoichiometry could be strongly
118 influenced by the climate and soil conditions. For example, edaphic properties

119 significantly affect the size, distribution density, and pattern of shrubs in arid and semi-
120 arid systems (Hughes et al., 2006), leading to changes in soil C:N. The accurate
121 evaluation and prediction of ecosystem structure and function require a comprehensive
122 investigation of the influence of shrub encroachment on soil C:N:P stoichiometry.

123 Here, we conducted a meta-analysis with 344 observations from 68 studies of
124 grasslands with a global distribution. The objectives of our meta-analysis were: (1) to
125 assess the responses of soil C, N, and P concentrations and stoichiometry to shrub
126 encroachment and (2) to explore whether such responses are mediated by shrub species,
127 soil factors, and climatic factors in shrub-encroached grasslands. We hypothesized that:
128 (1) shrub encroachment would increase soil C, N, and P concentrations and thus alter
129 element stoichiometry in grasslands; and (2) the responses of soil nutrient
130 concentrations and stoichiometry to shrub encroachment would depend on shrub
131 species identity and site environmental factors (i.e., mean annual precipitation, soil pH,
132 soil depth, and soil texture).

133

134 **Materials and methods**

135 **Data collection and extraction**

136 We set up a comprehensive database of responses of soil nutrient concentrations and
137 stoichiometry to woody plant encroachment from published global literature. First, we
138 searched the Web of Science and Google Scholar databases for online papers published
139 before July 2022. The combined search terms used in the database were '(either shrub,
140 bush, woody, shrubification, grassland, steppe, prairie, savanna, or shrubland) and
141 (either encroachment, expansion, thickening, proliferation, or colonization) and (soil
142 carbon, nitrogen, phosphorus, nutrient, stoichiometry, or C, N, P, C:N, C:P, N:P, C:N:P
143)' (Table S1 in Appendix S1). Then, we included those studies which satisfied the
144 following criteria: (1) only experiments conducted in grassland ecosystems were
145 included, i.e., those from tundra, forest, desert, and wetland were excluded; (2) shrub
146 encroachment was a natural process, without any artificial introduction, and field

147 experiments were conducted under natural conditions; (3) shrub-encroached grassland
148 was treated as one landscape type, and the variables were compared before and after
149 shrub encroachment (i.e., treatment and control); (4) sample size, mean values, and
150 standard deviations or errors of variables were reported.

151 Based on these criteria, we included 344 cases in our database, derived from 68
152 studies (Fig. 1, Table S2 in Appendix S1). Among the 344 independent plot-level
153 cases, 39% were from North America, 25% were from Asia, 11% were from Europe,
154 20% were from Africa, 3% were from South America and 2% were from Australia.
155 More than half of the studies were from savanna, while the rest were from semi-
156 arid grassland, arid grassland, and mesic grassland. All studies were from tropical
157 and mid latitudes. The latitudes ranged from 43°S to 46.56°N and the longitudes
158 from 110.88°W to 145.43°E; the mean annual precipitation was between 134 and
159 1559 mm, the mean annual temperature was between 1.1 and 24°C, and the average
160 altitude was between 75 and 3500 m.

161 For each publication, we collected the basic geographical information of the
162 experiment (site location, continent, latitude and longitude, landscape type,
163 precipitation, temperature, duration of shrub encroachment), the identity and basic traits
164 of the soil (pH, depth, texture), the taxonomic identity of the dominant encroaching
165 woody species (species, genus, and family), and the mean, standard deviation, and
166 sample size of the ecosystem responses that were assessed in plots with and without
167 woody plant encroachment. We extracted data on soil organic carbon (SOC), total
168 nitrogen (TN), and total phosphorus (TP) from the text or tables of each publication.
169 We extracted data from published figures using the GetData Graph Digitizer software
170 (version 2.26, <http://getdata-graph-digitizer.com>). In most publications, only the
171 response of element concentrations or element stocks to shrub encroachment was
172 reported. For our study, we calculated soil C:N, N:P, and C:P as molar ratios. For
173 publications where only soil organic matter concentration was reported, we calculated
174 soil total carbon (TC) concentration as soil organic matter concentration divided by a
175 constant of 1.724.

176

177 **Data analysis**

178 We defined the individual effect of shrub encroachment as the response of a specific
179 variable (e.g., C:P ratio) when compared with the control, that is, the grassland without
180 shrub encroachment, and we defined the grassland encroached by shrubs as the
181 experimental group. We evaluated the responses of soil C, N, and P concentrations and
182 their stoichiometry to shrub encroachment following the methods of Gurevitch and
183 Hedges (1999), with lnRR defined as the “effect size”. We conducted the meta-analysis
184 with the MetaWin 2.1 software (Rosenberg et al., 2000). For each pair of soil variables
185 from sites with and without encroachment, we calculated the response ratio of soil C,
186 N, and P concentrations and their stoichiometry as a measurement of effect size
187 (Rosenberg et al., 2000).

188
$$\ln RR = \ln\left(\frac{X_e}{X_c}\right) \quad (1)$$

189 where X_e and X_c are the means of variables from the experimental and control groups,
190 respectively. Positive and negative lnRR values indicate that shrub encroachment
191 increases and decreases element ratios compared with the control, respectively.

192 We estimated the variance (V) associated with lnRR from the standard deviation
193 of each mean value (Koricheva et al., 2013):

194
$$V = \frac{(SD_e)^2}{n_e X_e^2} + \frac{(SD_c)^2}{n_c X_c^2} \quad (2)$$

195 where SD_e and SD_c are the standard deviations of variables from the experimental and
196 control groups, respectively; X_e and X_c are the means of variables from the experimental
197 and control groups; and n_e and n_c are the sample sizes of variables from the experimental
198 and control groups. We calculated the standard deviation of nutrient concentrations
199 from the standard error if there only standard error was reported in a publication. We
200 estimated the standard deviation of element ratios based on similar experimental studies
201 (Gao et al., 2021).

202 We divided the soil into layers by depth: 0–20 cm depth was defined as the topsoil
203 layer, 20–60 cm as the middle soil layer, and 60–100 cm as the bottom soil layer. These
204 layers corresponded approximately to the A, B, and C horizons of each soil. We
205 classified the sites into four climate zones according to mean annual precipitation (MAP;

206 Middleton and Thomas, 1997): humid (> 800 mm), dry sub-humid (400–800 mm),
207 semi-arid (200–400 mm), and arid (0–200 mm). In addition, we classified the soil
208 samples into three pH categories: acidic (pH < 6.5), neutral (pH between 6.5 and 7.5),
209 and alkaline (pH > 7.5). Following the USDA taxonomy system (Brown, 1998; Staff,
210 1993), we further classified the soils in our database into nine texture types: sand, sandy
211 loam, loam, silty loam, sandy clay loam, clay loam, sandy clay, silty clay, and clay.
212 There were 23 shrub genera in our database: *Acacia*, *Caragana*, *Genista*, *Prosopis*,
213 *Schotia*, *Senegalia*, *Brachystegia*, *Chuquiraga*, *Cistus*, *Echinopartum*, *Eremophila*,
214 *Juniperus*, *Larrea*, *Morella*, *Nassella*, *Potentilla*, *Quercus*, *Rosmarinus*, *Artemisia*,
215 *Salix*, *Symphoricarpos*, *Spiraea*, and *Vaccinium*, which we classified into two groups:
216 Leguminosae and non-Leguminosae.

217 Several factors, such as climate (precipitation), soil properties (pH, depth, texture),
218 dominant plant (shrub genus, shrub families), and identity of encroachment (i.e., history
219 of shrub encroachment) were expected to influence the relationship between nutrient
220 status and shrub encroachment. We therefore considered these factors as moderator
221 variables in our meta-analysis.

222 We evaluated the overall effect of shrub encroachment on soil C, N, and P
223 concentrations and stoichiometry using a weighted random-effects model for meta-
224 analysis (Gurevitch and Hedges, 1999). We used Equation (3) to derive the weighted
225 mean response ratio ($\ln RR_{++}$).

$$226 \quad \ln RR_{++} = \frac{\sum_{i=1}^m \sum_{j=1}^k w_{ij} \ln RR_{ij}}{\sum_{i=1}^m \sum_{j=1}^k w_{ij}} \quad (3)$$

227 where $\ln RR_{++}$ is the weighted mean response ratio, m is the group number (e.g., soil
228 texture), k is the comparative number of i th group, and w is the weight of the response
229 ratio. Based on 999 iterations, we calculated the 95% confidence intervals (CI) of
230 $\ln RR_{++}$. We considered an experimental effect significant ($P < 0.05$) if its 95% CI did
231 not include zero (Koricheva et al., 2013).

232 We measured the total heterogeneity across studies using the Q_t statistic (Table S1
233 in Appendix S2). There was significant residual heterogeneity in the random effects for
234 the SOC concentration ($Q_t=1066.3$, $P < 0.0001$), TN concentration ($Q_t = 416$, $P <$

235 0.0001), and TP concentration ($Q_t = 252.0$, $P < 0.0001$) datasets, and for the C:N ($Q_t =$
236 292.0 , $P < 0.0001$), C:P ($Q_t = 381.7$, $P < 0.0001$), and N:P ($Q_t = 381.7$, $P < 0.0001$)
237 datasets, which we tried to explain with different moderators. We thus analyzed five
238 subgroups within the studies, i.e., climate classes, shrub families and species, soil layers,
239 soil textures, and soil pH categories.

240 The normal P-P plot (Fig. S2 in Appendix S2) showed that the effect sizes were
241 normally distributed. We assessed publication bias of each variable in our database with
242 funnel plots (Rosenberg et al., 2000). The funnel plots combined with the P values ($>$
243 0.05) based on an Egger's regression test showed no publication bias (Fig. S1 in
244 Appendix S2). The stability of an effect size increases and the size of the confidence
245 interval decreases with increasing sample size, which meant our large database yielded
246 accurate results. We tested the relationships among soil nutrient concentrations and
247 stoichiometric ratios using a Pearson correlation analysis. Moreover, we conducted a
248 linear regression analysis to explore the relationship between continuous variables
249 (duration of shrub encroachment) and changes in stoichiometric ratios. We did not
250 conduct this linear regression analysis when fewer than 30 studies were included. We
251 used SPSS 25.0 (SPSS Inc., Chicago, IL, USA) for both the Pearson correlation analysis
252 and the linear regression analysis.

253

254 **Results**

255 **Effects of climate and soil properties on soil C, N, and P concentrations and their** 256 **stoichiometry**

257 For the full database, the concentrations of SOC, TN, and TP increased significantly
258 after shrub encroachment, by 29%, 25%, and 20%, respectively ($P < 0.05$; Fig. 2a, b,
259 c). SOC concentration in the semi-arid, dry sub-humid, and humid zones increased by
260 33%, 36%, and 4%, respectively ($P < 0.05$), while no SOC response was detected in
261 the arid zone. SOC concentration were significantly enhanced in the topsoil layer, by
262 39%, after shrub encroachment. Furthermore, shrub encroachment increased SOC
263 concentration in clay, silty clay, sandy clay loam, sandy loam, and sand, by 36%, 40%,

264 32%, 36%, and 35%, respectively, but not in the other soil textures. SOC concentration
265 increased significantly, by 56%, with shrub encroachment in sites with neutral soil (Fig.
266 2a).

267 The response of soil TN concentration to shrub encroachment was only significant
268 in the dry sub-humid (+25%), semi-arid (+23%), and humid zones (+22%). Meanwhile,
269 TN concentration responded significantly to shrub encroachment in the topsoil and
270 middle soil layers, increasing by 29% and 13%, respectively, but did not show a
271 response in the bottom soil layer. TN concentration increased significantly with shrub
272 encroachment only in the clay, sandy loam, and sand soil textures, by 23%, 32%, and
273 29%, respectively. TN concentration increased significantly in all soil pH categories,
274 by 27% (acidic), 40% (neutral), and 17% (alkaline) (Fig. 2b).

275 TP concentration increased significantly in the humid (45%) and dry sub-humid
276 (21%) zones, but not in the semi-arid and arid zones. Similar to TN, TP concentration
277 increased significantly in the topsoil and middle soil layers, by 27% and 15%,
278 respectively, but did not respond significantly in the bottom soil layer. TP concentration
279 also increased significantly with shrub encroachment into clay, sandy loam, and sand,
280 by 128%, 13%, and 26%, respectively, but showed no significant responses in the other
281 soil texture s. The response of TP concentration to shrub encroachment was only
282 significant in acidic soil, where values increased by 32% (Fig. 2c).

283 Soil C:N, C:P, and N:P increased by 6%, 7%, and 7%, respectively, after shrub
284 encroachment ($P < 0.05$; Fig. 2d, e, f). Soil C:N increased significantly in the dry sub-
285 humid, semi-arid, and humid zones, by 6%, 3%, and 4%, respectively ($P < 0.05$), but
286 no significant response was detected in the arid zone. Shrub encroachment reduced soil
287 C:N by 11% in the bottom soil layer, but increased this ratio significantly in the topsoil
288 and middle soil layers ($P < 0.05$). The responses of soil C:N varied across soil textures,
289 with an enhancement in clay (18%), sandy clay loam (11%), loam (5%), and sandy
290 loam (4%) but no significant response in sandy clay, clay loam, and sand. Soil C:N
291 increased significantly in acidic and neutral soils, but there was no significant response
292 in alkaline soil (Fig. 2d).

293 Soil C:P only increased significantly with shrub encroachment in the dry sub-

294 humid zone (by 12%). In addition, soil C:P increased in the topsoil layer (by 19%) and
295 in sandy loam (by 29%) after shrub encroachment, but no significant response was
296 detected in other soil layers or soil textures. Additionally, soil C:P increased
297 significantly in acidic and neutral soils, by 18% and 29%, respectively, but no
298 significant response was observed in alkaline soil (Fig. 2e).

299 Shrub encroachment reduced soil N:P in the humid zone, by 33%, but increased
300 this ratio in the dry sub-humid (8%) and semi-arid (30%) zones. Soil N:P only
301 responded significantly in the topsoil layer, with an increase of 12%. In addition, shrub
302 encroachment increased soil N:P significantly (by 23%) in sandy loam but decreased
303 significantly (by 34%) in clay, while no response was detected in other soil textures.
304 Shrub encroachment enhanced soil N:P in acidic and neutral soils, by 7% and 16%,
305 respectively, while no significant effect was observed in alkaline soil (Fig. 2f).

306

307 **Effects of shrub identity on soil C, N, and P concentrations and their stoichiometry**

308 SOC concentration increased significantly when encroached by Leguminosae and non-
309 Leguminosae, by 39% and 30%, respectively. Regarding leguminous shrubs, SOC
310 concentration increased significantly with encroachment by *Caragana*, *Genista*, and
311 *Prosopis*, by 17%, 39%, and 52%, respectively (Fig. 3a). For non-leguminous shrubs,
312 SOC concentration increased significantly with encroachment by the genera of *Morella*
313 (+140%), *Potentilla* (+36%), *Rosmarinus* (+44%), *Artemisia* (+100%), and *Vaccinium*
314 (+166%; Fig. 3a). TN concentration were enhanced significantly when encroached by
315 Leguminosae and non-Leguminosae, by 25% and 18%, respectively. Specifically, TN
316 concentration increased significantly when encroached by *Acacia*, *Prosopis*, *Morella*,
317 *Potentilla*, *Quercus*, and *Vaccinium*, by 24%, 34%, 133%, 18%, 24%, and 94%,
318 respectively (Fig. 3b). Overall, TP concentration increased significantly when
319 encroached by Leguminosae and non-Leguminosae, by 16% and 35%, respectively. TP
320 concentration only increased significantly (by 128%) when the encroaching genus was
321 *Vaccinium* (Fig. 3c).

322 Overall, soil C:N increased significantly with encroachment by Leguminosae and non-
323 leguminosae, by 4% and 6%, respectively. Specifically, when encroached by the

324 leguminous genera *Caragana* and *Prosopis*, C:N increased by 4%. Soil C:N also
325 increased significantly when encroached by the non-leguminous genera of
326 *Brachystegia* (+15%), *Potentilla* (+7%), *Quercus* (+4%), *Symphoricarpos* (+13%), and
327 *Spiraea* (+9%; Fig. 3d). Non-leguminous shrubs enhanced soil C:P by 31% overall,
328 while leguminous shrubs did not affect this ratio. Soil C:P decreased significantly, by
329 18%, with encroachment by the leguminous *Caragana* genus. In contrast, soil C:P
330 increased significantly with *Prosopis* (leguminous) and *Potentilla* (non-leguminous)
331 encroachment, by 22% and 34%, respectively (Fig. 3e). Overall, soil N:P increased
332 significantly, by 13%, with encroachment by non-leguminous shrubs, but not by
333 leguminous shrubs. Regarding individual genera, soil N:P decreased with the
334 encroachment of *Caragana* (leguminous) and of *Echinopartum*, and *Vaccinium* (non-
335 leguminous), by 20%, 32%, and 34%, respectively. Furthermore, soil N:P increased by
336 24%, 18%, and 43% when encroached by the genera *Prosopis*, *Potentilla*, and *Salix*,
337 respectively, but it showed no significant change with the other encroaching shrub
338 genera (Fig. 3f).

339

340 **Soil C, N, and P concentrations and their stoichiometry in relation to shrub** 341 **encroachment duration**

342 The changes in SOC and TN concentrations were positively correlated with the duration
343 of shrub encroachment (both $P < 0.001$; Fig. 4a, b), whereas the changes in soil C:N
344 were negatively correlated with the duration of shrub encroachment ($P = 0.0017$; Fig.
345 4c). There were negative correlations between soil P concentration and C:P, between P
346 concentration and N:P, and between C:N and N:P (all $P < 0.01$; Table S2 in Appendix
347 S2). SOC concentration was positively correlated with P concentration, C:N, and C:P
348 (all $P < 0.01$). Similarly, soil N concentration was positively correlated with P
349 concentration ($P < 0.01$), C:N ($P < 0.05$), C:P ($P < 0.01$), and N:P (all $P < 0.01$).
350 Additionally, there were negative correlations between soil P concentration and C:P
351 and N:P, and a positive correlation between soil P concentration and C:N (all $P < 0.01$).

352

353 **Discussion**

354 **Responses of soil nutrient concentrations and stoichiometry to shrub** 355 **encroachment**

356 SOC, TN, and TP concentrations and their stoichiometric ratios changed in response to
357 shrub encroachment. There are at least five possible reasons for such responses. First,
358 shrubs are less nutrient conservative than herbs (Elser et al., 2000; Thompson et al.,
359 1997), leading to greater C and N storage. In particular, shrub litter is more recalcitrant
360 than grass litter (Montané et al., 2010) due to higher concentrations of C-rich
361 recalcitrant biopolymers. For instance, the litter and roots of woody plants have high
362 levels of cinnamyl and other compounds that are resistant to mineralized, thus being
363 harder to decompose (Opsahl and Benner, 1997). Second, higher litter quantities and
364 NPP in shrub-encroached grasslands (Ding et al., 2019) can increase SOC storage
365 (Boutton et al., 2009). Third, shrub encroachment enhances the soil retention capacity
366 and reduces leaching and runoff, which can improve soil P retention (Ding et al., 2019).
367 Fourth, shrubs with deeper root systems deposit more C in the soil (Wang et al., 2016)
368 than occurs in grasslands without shrubs. Fifth, the fixation of N by leguminous shrub
369 can enhance soil N concentration (Su and Zhao, 2003). Shrub encroachment is
370 conducive to the sequestration of C and N, but the disproportionate increases in C and
371 N may lead to an imbalance in soil C:N. In addition, the relationship between soil C:N
372 stoichiometry and microbial activity may affect C sequestration, as microbes have a
373 lower C use efficiency when the C:N ratio is high (> 15) (Alberti et al., 2014). Increases
374 in soil C:N after shrub encroachment may additionally enhance soil N availability
375 (Alberti et al., 2014).

376

377 **The role of climatic factors**

378 The increases in soil SOC and N with shrub encroachment reached a maximum in dry
379 sub-humid and semi-arid climate zones, suggesting that medium precipitation levels are
380 good for C sequestration during shrub encroachment. Simultaneously, soil C:N and N:P

381 increased significantly in dry sub-humid and semi-arid grasslands, and C:P also
382 increased significantly in the dry sub-humid climate zone. Soil C:N:P stoichiometry is
383 strongly effected by soil moisture, which is altered strongly by precipitation intensity
384 (Sun et al., 2020; Yuan and Chen, 2015). Previous studies have indicated that soil C
385 and N concentrations in shrub-encroached grasslands may increase (Wheeler et al.,
386 2007) or decrease (Jackson et al., 2002) with increasing precipitation. Changes in
387 precipitation could significantly influence the wetting-drying cycle of soil, litter mass
388 (Denef et al., 2001), and the activities of microorganisms (Xu et al., 2021). Soil
389 microbial activity increases considerably with high annual precipitation in general,
390 which can have a significant effect on litter decomposition rate, thus resulting in
391 changes in the use of soil nutrients (Denef et al., 2001) and in the nutrient limitation
392 situation. In our study, soil N:P increased significantly with shrub encroachment in the
393 semi-arid zone, as the increase in N far exceeded the increase in P, which may have
394 resulted in a P limitation in the soil. In contrast, we found that the changes in soil C:N
395 after shrub encroachment were similar along the precipitation gradient, indicating that
396 the response of soil C:N to shrub encroachment is less sensitive to variation in climate.

397

398 **The role of soil conditions**

399 The changes in C, N, and P concentrations were much stronger in the topsoil layer than
400 in deeper layers, which is similar to the finding of Boutton et al. (2009). This may be
401 because the SOC in topsoil mainly comes from leaf litter and root exudates (Wu et al.,
402 2023), and the greater amount of litter input into the topsoil after shrub encroachment
403 (as shrubs have a greater biomass) enhances C accumulation in the topsoil. The roots
404 of shrubs are mainly distributed in the 0–60 cm depth soil layer, so the bottom soil layer
405 (60–100 cm depth) receives less leaf litter and root exudates compared with the topsoil,
406 explaining why there was no significant increase in SOC in the bottom soil layer. Soil
407 C:N, C:P, and N:P increased significantly in the topsoil because the increases in SOC
408 and TN concentrations in the topsoil were greater than that of TP. This may be caused
409 by the relatively simple input path of P (Schlesinger and Bernhardt, 2020). Our results

410 demonstrate that there exists a strong edaphic control of soil nutrient element
411 accumulation and soil stoichiometry associated with shrub encroachment. Soil texture
412 has been reported to significantly affect the woody-to-grass ratio, due to its effects on
413 soil water content, plant growth, and nutrient concentrations and availability (Britz and
414 Ward, 2007). We found that the concentrations of SOC, TN, and TP, as well as C:N,
415 increased with shrub encroachment in clay soil. Physical protection of SOC varies with
416 soil structure. Liao et al. (2006) reported that about 30–55% of the SOC generated after
417 shrub encroachment was protected within macro- and microaggregates or correlated
418 with the fraction of silt and clay. Fine-textured soil (higher silt and clay contents) is
419 characterized by organomineral complexes and the formation of macro- and
420 microaggregates, which protect organic matter from mineralization and/or eluviation
421 (Campbell et al., 1998) and from decomposition by soil microbes (Ladd et al., 1993;
422 Christensen, 2020). In our meta-analysis, soil nutrient element concentrations increased
423 significantly after shrub encroachment in sandy soil, which may be because coarse-
424 textured soils tend to have a higher bulk density (Håkansson and Lipiec, 2000).
425 Meanwhile, soil N:P responded negatively in clay soil, which may be attributed to the
426 fact that the magnitude of TP increase far exceeded the magnitude of TN increase.
427 Acidic soil affects the decomposition rate of litter and the release rate of soil C and N
428 (Hättenschwiler and Bretscher, 2010). Moreover, the C storage in acidic soil is greater
429 under high N conditions than under low N conditions (Hagedorn et al., 2003).
430 Consistent with these previous findings, we observed that SOC and TN increased in
431 acidic soil after shrub encroachment, especially by Leguminosae shrubs. In neutral soil,
432 C:N, C:P, and N:P increased significantly, which may be attributed to the large
433 increases in TN and especially SOC concentration after shrub encroachment in neutral
434 soil, combined with the smaller change in TP concentration (Fig. 2a, b, c).

435

436 **The role of woody plant type**

437 Shrub type could affect soil nutrient concentrations by altering litter quality and nutrient
438 uptake. The spread of N-fixing shrubs leads to massive reactive nitrogen enrichment
439 (Bühlmann et al., 2014) and may be an important driving factor in the increases in SOC

440 observed here. The encroachment of non-leguminous shrubs led to a significant
441 increase in C:P in our meta-analysis, reflecting the lower availability of P (Wang and
442 Yu, 2008). A greater increase in soil C:P in grasslands encroached by non-leguminous
443 shrubs therefore results in a decrease in available P that can be absorbed by plants for
444 growth. Leguminous shrub encroachment did not affect soil C:P significantly,
445 reflecting a non-significant effect on available P.

446

447 **Conclusions**

448 In summary, our global synthesis verifies that shrub encroachment promotes soil C
449 sequestration and increases soil C:N, C:P, and N:P. Moreover, SOC and TN accumulate
450 over time after shrub encroachment. Most importantly, shrub encroachment
451 simultaneously shifts soil stoichiometry in grasslands with sandy loam, acidic, and
452 neutral soils, as well as in those with a dry sub-humid climate. The rate of increase in
453 soil C:N decreases with the duration of shrub encroachment, suggesting a stabilization
454 of the soil C:N ratio in shrub-encroached grasslands. These findings provide
455 fundamental knowledge for estimating the impacts of shrub encroachment (partly due
456 to global climate change) on C and N cycling in grasslands, and they help to better
457 predict the effects of global change on grassland biogeochemical cycling and ecosystem
458 services in a changing world.

459

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472

473 **Conflicts of interest**

474 The authors declare that they have no known competing financial interests or personal
475 relationships that could have appeared to influence the work reported in this paper.

476

477 **Author contributions**

478 This study was conceived by Z.D. and H.Z. The data was collected and analyzed by
479 H.Z., D.D., X.C., D.G., Y.H., and S.N. The manuscript was written by Z.D. and
480 H.Z., with input from J.P., J.S., X.L. and M.L. and comments from all other authors.

481

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