

This is the **accepted version** of the journal article:

Wang, Qin; Peñuelas, Josep; Tan, Bo; [et al.]. «The role of decaying logs in nursing soil fungal diversity varies with decay classes in the forest ecosystem». European Journal of Soil Science, Vol. 73, issue 3 (May–June 2022), e13243. DOI 10.1111/ejss.13243

This version is available at <https://ddd.uab.cat/record/299909>

under the terms of the  ^{IN}COPYRIGHT license

1 **Title:** The role of decaying logs in nursing soil fungal diversity varies with decay
2 classes in the forest ecosystem

3 **Running head:** Soil fungi vary with log decay classes

4 **Authors:** Qin Wang^{a,b,c,d}, Josep Peñuelas^{c,d}, Bo Tan^b, Zhuang Wang^b, Han Li^b, Rui
5 Cao^b, Chenhui Chang^{b,e}, Yurui Jiang^b, Wanqin Yang^{a*}

6 **Affiliations:** ^aSchool of Life Science, Taizhou University, Taizhou 318000, Zhejiang,
7 PR China

8 ^bInstitute of Ecology and Forestry, Sichuan Agricultural University, 211 Huimin Road,
9 Chengdu 611130, PR China

10 ^cCSIC, Global Ecology Unit CREAM-CSIC-UAB, CSIC, Bellaterra 08193, Barcelona,
11 Catalonia, Spain

12 ^dCREAF, Cerdanyola del Vallès 08193, Barcelona, Catalonia, Spain

13 ^eCollege of Urban and Environmental Sciences, Peking University, Beijing 100871, PR
14 China

15 ***Corresponding author:**

16 Wanqin Yang

17 Postal address: 1139# Shifu Avenue, Taizhou 318000, Zhejiang, China

18 E-mail: scyangwq@163.com

19 Tel: 86-576-88660339, Fax: 86-576-88660339

20

21 **Abstract**

22 Soil fungi are crucial drivers of log decomposition in forest ecosystems, but how soil
23 fungal community composition varies during the process of log decomposition remains
24 poorly understood. We conducted an experiment incubating decaying logs in a
25 subalpine coniferous forest on the eastern Qinghai-Tibet Plateau, China. Five classes of
26 decaying Minjiang fir (*Abies faxoniana*) logs were incubated on the forest floor, and
27 the composition and diversity of fungal communities in soils underneath decaying logs
28 were measured using high-throughput sequencing. A total of 4321 operational
29 taxonomic units (OTUs) were detected by Illumina Novaseq sequencing analysis. Soil
30 fungal diversity differed significantly during the process of log decomposition and was
31 highest in decay classes III or IV. Basidiomycota and Ascomycota were dominant phyla
32 regardless of log decay classes. Moreover, the proportion of arbuscular mycorrhiza,
33 wood saprotroph and saprotroph increased during the process of log decomposition, but
34 that of ectomycorrhiza decreased. The structure of soil fungal community underneath
35 decaying logs varied greatly with decay classes. Different decay classes of logs favor
36 special fungal groups, implying that the ecological effects of logs at differing decay
37 classes on soil fungal communities were different.

38 **Key words:** *Abies faxoniana*, coarse woody debris, community composition, functional
39 group, soil fungal, subalpine forest

40 **1.Introduction**

41 Decaying logs are an important habitat for many organisms and play crucial roles in

42 the global carbon cycle, soil formation, and water and soil conservation (Harmon et al.
43 1986; Lassauce et al. 2011; Forrester et al. 2012). During the process of log
44 decomposition, fungi exhibit strong priority effects (Fukami et al. 2010, Leopold et al.
45 2017). Fungal hyphae can penetrate into the woody debris and secrete extracellular
46 enzymes to decompose refractory organics such as lignocellulose (Baldrian and Banin
47 2017), finally releasing soluble substrates for microbial uptake (Moorhead et al. 2012).
48 The logs are in direct contact with soil surface and therefore might alter the fungal
49 community structure of underlying soil during the decomposition process. Zalamea et
50 al. (2016) and Semchenko et al. (2018) have reported that the decomposition of logs
51 directly affects the physical, chemical and biological properties of soils underneath the
52 logs. Minnich et al. (2021) have also found that deadwood can affect soil microbial
53 community structure. Previous studies have reported the role of deadwood in
54 conservation of macrofungi, plants and invertebrates (Bani et al. 2018), but the
55 responses of composition and functional groups of soil fungal community to log
56 decomposition have rarely been studied.

57 Soil fungi are key participants in the biogeochemical cycles of forest ecosystems,
58 playing essential roles in the formation and development of soil, storage of soil carbon
59 (C), nutrient cycling and energy flow (Tedersoo et al. 2014). The role of soil fungal
60 community is closely related to its structure and diversity, and the structure and
61 diversity of the soil fungal community are affected by biological and abiotic factors
62 such as soil moisture and temperature and substrate quality (Chen et al. 2017b;
63 Asemaninejad et al. 2018). Lindahl et al. (2007) have observed the presence of

64 saprotrophic, wood-decomposer fungi in soils. These fungi are main factors in
65 decomposing organic matter and are critical in regulating nutrient cycling. Some fungal
66 species are found solely in the soil underneath decaying wood (Mäkipää et al. 2017).
67 Some mycorrhizal fungi have ability to colonize the wood substrate in contact with soil
68 surface, break down organic matter, and even tend to produce fruiting bodies more often
69 in wood than in soil (Lindahl and Tunlid 2015). In addition, previous studies have
70 observed that the effect of deadwood on soil varies with the decay stage (Bł óńska et al.
71 2017; Zalamea et al. 2007), and that the stage of decomposition has an impact on the
72 C:N:P stoichiometry of the soil (Wojciech et al. 2019). There is no obvious sign of
73 decomposition in newly formed logs (Ganjegunte et al. 2004), while highly decayed
74 logs have a larger contact area with the forest floor, thus favouring sediment adsorption
75 (Barker, 2008), and releasing more ions into the surface layers of soil than less-
76 decomposed wood, which may further cause differences in soil fungal community
77 structure (Jarosław et al. 2017). Thereby, the composition and diversity of soil fungal
78 community might be closely related with the log decay process.

79 Because it is affected by the limitation of low temperature and frequent geological
80 disasters, such as earthquakes and mudslides, the soil in the subalpine forest on the
81 eastern Qinghai-Tibet Plateau develops slowly (Yang and Wu 2021). Therefore, soil
82 organic layer and the decaying logs on the ground play an important role for
83 biodiversity conservation, nutrient cycling, and water and soil conservation (Yang and
84 Wu 2021). The goal of this study was to analyze the effects of decaying logs on the
85 composition of soil fungal communities and the response of the fungal community and

86 functional groups to the decay classes. We hypothesized that (1) decaying logs favoured
87 different groups of soil fungi thus altering soil fungal communities; and (2) highly
88 decayed logs would lead to stronger effects on soil fungal communities. We tested these
89 hypotheses by incubating decaying logs in a subalpine coniferous forest on the eastern
90 Qinghai-Tibet Plateau, China. Minjiang fir (*Abies faxoniana*) is the dominant tree
91 species in this region. Minjiang fir logs at various stages of decay were incubated for
92 six years on the forest floor, and the compositions of the soil fungal communities in the
93 soil underneath the logs were determined using high-throughput sequencing. The
94 objective of this study was thus to assess the ecological effects of logs at differing decay
95 classes on soil fungal communities.

96 **2. Materials and methods**

97 **2.1 Site description**

98 The study was conducted at the Long-Term Research Station of Alpine Forest
99 Ecosystems (31°14'-31°19'N, 102°53'-102°57'E; 2458-4619 m a.s.l.) in Li County,
100 Sichuan Province, southwestern China. The region is a transitional area between the
101 Tibetan Plateau and the Sichuan Basin. The annual mean temperature and precipitation
102 are approximately 2-4 °C and 850 mm, respectively (Chang et al. 2019). The seasonal
103 soil freeze-thaw period begins in early November and ends in April the following year.
104 The soil is classified as Cambisols. The accumulation of large amounts of woody debris
105 (53 Mg·ha⁻¹) in this region (Xiao et al. 2016) plays crucial roles in maintaining soil
106 fertility, controlling soil erosion and promoting biodiversity. The trees consist mainly
107 of Minjiang fir (*Abies faxoniana*), cypress (*Sabina saltuaria*), larch (*Larix mastersiana*)

108 and red birch (*Betula albo-sinensis*), with Minjiang fir accounting for 80% of the total
109 basal area at the site (Xiao et al. 2014). The main understory shrubs include willow
110 (*Salix paraplesia*) and azalea (*Rhododendron lapponicum*).

111 **2.2 Experimental design**

112 We established a permanent 100 × 100 m plot of primary Minjiang fir forest in the
113 nature reserve in August 2013 based on previous studies (Chang et al. 2019; Chang et
114 al. 2020) and three 25 × 25 m subplots with similar slope aspects, slope degrees, canopy
115 closures and tree heights. Logs of Minjiang fir with 35 ± 5 cm in diameter in five decay
116 classes (I-V, at increasing levels of decay) described in detail by Chang et al. (2019)
117 and Wang et al. (2021) were collected. The logs of decay class I were from newly felled
118 trees, representing the start of decomposition, and the logs of classes II-V were
119 collected from the primary Minjiang fir forest neighbored the permanent plot. All logs
120 were sawed on-site into lengths of 120 cm and carefully placed on the forest floor under
121 the closed canopy in each subplot. See Chang et al. (2020) and Wang et al. (2021) for
122 more details about the placement and arrangement of decaying logs.

123 **2.3 Sampling**

124 In August 2019, after decaying logs were carefully removed, soil samples were
125 collected in each subplot using an auger (10 cm long, 5 cm diameter) from the organic
126 layer of the forest floor. Control samples (CK) were also collected 3 m away from the
127 decaying logs. The samples were transported to the laboratory on ice packs in a cooler.
128 A total of 54 soil samples were collected, i.e., (1 CK + 5 decay classes) × 3 replicates

129 × 3 subplots. Visible living plant material was removed from the samples, and about
130 20g of soil was used to determine the water content (WC). The samples were passed
131 through a 2-mm sieve, about 10g of air-dried soil samples were used for pH
132 determination, and subsamples were stored at -70 and 4 °C for molecular and
133 biochemical analysis, respectively. Five of the nine replicate samples were randomly
134 selected for molecular biochemical analysis, thus making a total of 30 samples.

135 **2.4 Soil physiochemical and biological analyses**

136 Water content was determined by weighing the samples before and after drying. Soil
137 pH was measured using a pH meter at a soil: water ratio of 1:2.5 (m/v). The
138 concentrations of organic C (OC), total nitrogen (TN) and total phosphorus (TP) in soil
139 were measured using potassium dichromate oxidation-ferrous sulfate titrimetry,
140 Kjeldahl determination and molybdenum- antimony colorimetry, respectively (Lu
141 1999). Soil microbial biomass C (MBC) and nitrogen (MBN) were measured using
142 chloroform fumigation and K₂SO₄ extraction (Brookes et al. 1985; Vance et al. 1987).
143 Soil microbial biomass phosphorus (MBP) was measured using chloroform fumigation
144 and NaHCO₃ extraction (Brookes et al. 1982).

145 **2.5 DNA extraction, PCR amplification and high-throughput sequencing**

146 Microbial DNA was extracted from the samples using the FastDNA Spin Kit for Soil
147 (MP Biomedicals, Solon, USA) following the manufacturer's instructions. DNA was
148 checked for quality using agarose gel electrophoresis and ultraviolet absorbance
149 (ND1000, Nanodrop, Thermo Fisher Scientific, Waltham Mass, USA). The DNA was
150 diluted to 1 ng/μl using sterile water and stored at -20 °C until analysis.

151 A region of the fungal ITS2 gene was amplified, purified, quantified, pooled and
152 sequenced on an Illumina NovaSeq platform at Novogene Genomics Institute (Beijing,
153 China). The fungal ITS2 gene was amplified using the primer pair ITS3-2024F
154 (GCATCGATGAAGAACGCAGC)/ITS4-2409R (TCCTCCGCTTATTGATATGC).
155 All the samples were amplified in 3 replicates, and no-template controls were included
156 in all steps of the process. Triplicate PCR amplicons were pooled together, after which
157 they were detected by electrophoresis in a 2% (w/v) agarose gel, and bands were
158 excised, mixed in equal density ratios and purified using a GeneJET Gel Extraction Kit
159 (Thermo Fisher Scientific, Waltham Mass, USA). The purified PCR products were
160 sequenced on the Illumina NovaSeq platform by Pair-End 250 The barcode sequences
161 were spliced with FLASH (v 1.2.7) software to obtain the raw tags (Zhang et al. 2014).
162 The clean tags were obtained by quality filtering the raw tags using Quantitative
163 Insights Into Microbial Ecology (QIIME) (v1.9.0) software (Caporaso et al. 2010;
164 Bokulich et al. 2013). Chimeric sequences were removed using the UCHIME algorithm
165 (Edgar et al. 2011). USEARCH was used to identify and remove chimeras, and the
166 remaining sequences were clustered to generate operational taxonomic units (OTUs) at
167 the 97% similarity level (Edgar 2010; Magoč and Salzberg 2011). The UNIT (v 7.2)
168 database (Kõljalg et al., 2013) and QIIME (v 1.9.0) (Altschul, 1990) software were
169 selected as species annotation analysis for ITS. A rapid multiple sequence alignment
170 was performed using MUSCLE (Edgar, 2004) software to obtain the phylogenetic
171 relationship of all OTUs representing sequences. Normalization was performed based
172 on the least amount of data in each sample, and subsequent analyses were based on the

173 normalized data.

174 In addition to fungal community composition, functional groups were assigned based
175 on FUNGuild (Nguyen et al. 2016). For analysis, we classified the fungal taxa into six
176 functional groups (i.e., animal pathogen, arbuscular mycorrhiza, ectomycorrhiza, plant
177 pathogen, wood saprotroph and saprotroph) with some overlaps. ‘other’ (including
178 ericoid mycorrhizal fungi, foliar endophytes, lichenized fungi, mycoparasites, etc.), and
179 undetermined (OTUs that were not identified to genus or whose genus was not
180 represented in the FUNGuild database) were included in analysis but the results on
181 guild level were not shown. We considered only FUNGuild assignments with a
182 confidence score of ‘probable’ or ‘highly probable’ and classified taxa with assignment
183 scores below these as undetermined (Lankau and Keymer. 2016).

184 **2.6 Statistical analyses**

185 The Shannon and Simpson diversity indices, metrics of alpha diversity of the soil
186 fungal communities, were calculated. The significance of the differences in alpha
187 diversity of soil fungi among log decay classes was tested by Tukey's honestly
188 significant difference (HSD) test when the results of variance analysis conducted with
189 general linear models were significant at $P = 0.05$. One-way ANOVA and HSD test
190 were also performed to test the effects of decay class on the abundance of the fungal
191 community. The effects of log decay classes on soil fungal communities were tested
192 using permutational multivariate analysis of variance (PERMANOVA) based on Bray-
193 Curtis distance and the analysis of similarity (ANOSIM), and visualised using two-
194 dimensional Non-metric multidimensional scaling (NMDS) analysis based on Bray-

195 Curtis distance. Heatmap was used to show the relative abundance of dominant genera
196 and the proportion of functional groups of soil fungi from different decay classes of
197 logs. Spearman correlation analysis was used to identify relationships between soil
198 physicochemical properties and the relative abundances of fungal taxa. One-way
199 ANOVA was performed using SPSS 20.0 (IBM Corporation, Armonk, USA). Other
200 statistical analyses were performed using the VEGAN package in R (Oksanen et al.
201 2013; R Development Core Team 2015). The statistical tests were considered
202 significant at $P < 0.05$.

203 **3. Results**

204 **3.1 Characteristics of the sequence data and species alpha diversity**

205 Across 30 soil samples underneath decaying logs, a total of 3 344 637 raw fungal
206 sequences, and 1 926 264 high-quality fungal sequences were retained after the raw
207 sequences were controlled for quality. The sequence data were classified into 4321
208 fungal OTUs at a cutoff in sequence similarity of 97%. The Shannon and Simpson
209 diversity indices for the fungal communities differed significantly among the decay
210 classes ($P < 0.05$) and were the highest for decay classes III and IV (Table 1).
211 However, the number of OTUs of soil fungal community was not significantly different
212 among log decay classes and neither between CK and presence of logs ($P > 0.05$) (Table
213 1). Compared with CK, soil fungi communities under decaying logs had more
214 independent OTUs (Fig. S1a), which was most for decay classes II and V (Fig. S1b).

215 3.2 Composition of fungi

216 Basidiomycota (22.0-49.3%), Ascomycota (15.7-41.5%) and Mortierellomycota
217 (2.4-11.9%) were the most dominant fungal phyla, representing >65% of the total reads
218 (Fig. 1a). The relative abundance of Mortierellomycota was significantly higher in the
219 soil underneath decaying logs than in CK and was the lowest in decay class II and the
220 highest in decay class IV ($P < 0.05$), while the relative abundances of Basidiomycota
221 and Ascomycota did not differ significantly among the classes ($P > 0.05$) (Figs. 1a and
222 S2a). The relative abundances of other phyla, including Rozellomycota,
223 Glomeromycota, Chytridiomycota and Mucoromycota, were less than 2%. The relative
224 abundances of these phyla were significantly affected by the decaying logs except
225 Mucoromycota.

226 A total of 24 fungal genera (with relative abundance >1%) were identified across all
227 samples (Fig. 1b). *Piloderma*, *Inocybe* and *Mortierella* were the three most abundant
228 genera. The relative abundance of *Piloderma* in soils under decaying logs for decay
229 classes I, II and IV was significantly higher than in CK, and significantly lower for
230 decay classes III-V than I-II. The relative abundance of *Inocybe* in soils under decaying
231 logs for decay classes I, II, IV and V was lower than in CK, and were lowest for decay
232 class V. The relative abundance of *Mortierella* in soils under decaying logs for all decay
233 classes was significantly higher than in CK (Figs. 1b and S2b). The relative abundances
234 of most of the other genera, such as *Ilyonectria*, *Fusarium*, *Phlebia*, *Humicola* and
235 *Pleotrichocladium*, were significantly higher in general under the logs than in CK (Fig.
236 S2b).

237 Based on Bray-Curtis, PERMANOVA, ANOSIM and NMDS were used to compare
238 the difference in composition of soil fungi between decay classes (Fig. 2). The
239 composition of the soil fungal communities differed significantly among the classes
240 (stress <0.2, ANOSIM, R = 0.628, P = 0.001). Additionally, PERMANOVA test verified
241 that soil fungal communities significantly differed among log decay classes and were
242 also affected by presence of logs (Table S1).

243 **3.3 Compositions of fungal functional group inferred by FUNGuild**

244 The relative abundance of most putative fungal functional groups was significantly
245 affected across the treatments (Fig. 3). The proportion of arbuscular mycorrhiza, wood
246 saprotrophs and saprotroph increased with the increase of decay classes of logs, while
247 the proportion of ectomycorrhiza decreased, while the relative abundance of wood
248 saprotrophs did not differ between CK and decay class I - III. The relative abundance
249 of animal pathogen was highest for decay class IV, while the relative abundance of plant
250 pathogen was highest for decay class III. Additionally, heatmap made based on fungal
251 community composition and functional guilds showed that the community composition
252 of soil fungi was similar under decay classes IV and V of logs, but different from that
253 of CK, while those of soil fungi communities under early decomposition stage were
254 more similar to that of CK (Fig. 4a and Fig. 4b).

255 **3.4 Correlations between fungal composition and soil physicochemical properties**

256 Spearman correlation heatmaps were analyzed to identify the relationships of the
257 fungal communities with biochemical properties and environmental factors (Fig. 5).

258 Soil pH, microbial biomass carbon, microbial biomass nitrogen and total phosphorus
259 were important factors correlated to soil fungal community composition. The relative
260 abundance of Basidiomycota was correlated positively with microbial biomass carbon,
261 microbial biomass nitrogen and total phosphorus, while the opposite trend was found
262 for Ascomycota, Glomeromycota and Chytridiomycota (Fig. 5a). Microbial biomass
263 carbon, microbial biomass nitrogen, total phosphorus and water content were
264 negatively correlated with the relative abundances of the genera *Ilyonectria*, *Fusarium*,
265 *Humicola*, *Pleotrichocladium*, *Chaetomium*, *Tetracladium*, *Dactylonectria*, and
266 *Didymosphaeria*, but microbial biomass carbon, microbial biomass nitrogen, microbial
267 biomass phosphorus and organic carbon were positively correlated with the relative
268 abundances of *Cortinarius* and *Meliniomyces*. Soil pH was positively correlated with
269 the relative abundances of *Ilyonectria*, *Fusarium*, *Humicola*, *Pleotrichocladium*,
270 *Chaetomium*, *Tetracladium* and *Dactylonectria* (Fig. 5b).

271 **4. Discussion**

272 Level of decay in logs affects soil fungal community, consistent with our hypotheses.
273 The diversity of soil fungal community was highest in decay classes III or IV (Table 1),
274 and it tended to stabilize after that. Changes in fungal community were mostly
275 explained by changes in genera that were not detected on phylum level. Additionally,
276 the proportion of functional groups varied with decay classes of logs (Fig. 4), indicating
277 that decaying logs favor different groups of soil fungi thus altering the composition of
278 soil fungal communities. The composition of soil fungal communities underneath

279 decaying logs was closely related to physicochemical properties (e.g., microbial
280 biomass nitrogen, total phosphorus concentrations and water content. Our results
281 indicated the ecological effects of logs at differing decay classes on soil fungal
282 community were different.

283 **4.1 The diversity of soil fungal communities differed significantly among log decay** 284 **classes**

285 The Shannon and Simpson diversity of fungal communities in soils underneath
286 decaying logs increased during the decaying process, while it tended to stabilize after
287 that. This is likely that decaying logs provide an additional source of nutrients, and
288 some microbial species, such as fungi utilizing wood, occur in soil only under decaying
289 logs (Mäkipää et al. 2017). Additionally, highly decayed logs provide a more stable
290 moisture content of the soil surface (Zalamea et al. 2016), offering beneficial habitat
291 conditions for some fungal species. However, logs in decay class V were basically
292 fragmented material, with more refractory substances, leading to the replacement of
293 some fungi, so the fungal diversity of soils underneath decay class V was lower than
294 underneath decay class IV and then tended to be stable.

295 **4.2 Effect of decaying logs on composition of soil fungi**

296 Basidiomycota and Ascomycota were the two dominant phyla across CK and all
297 decay classes, consistent with previous studies (Chen et al. 2017a; Zhang et al. 2021).
298 Some species of Basidiomycota and Ascomycota have the ability to decompose
299 refractory substances such as lignin and cellulose, but the existence of so many species

300 in these two phyla, lead to the absence of any significant difference in the relative
301 abundance of these phyla between decay classes. Mortierellomycota, a typical phylum,
302 is mostly saprophytic in soil (Lindahl et al. 2007), and its relative abundance was
303 significantly different in soil among decay classes, which is consistent with a previous
304 study (Mäkipää et al. 2017). The significant differences in microbial biomass nitrogen
305 and total phosphorus concentrations among decay classes may to some extent account
306 for the separation of Mortierellomycota.

307 *Piloderma* and *Cortinarius* declined in abundance during the process of log
308 decomposition, while *Inocybe*, *Russula* and *Lactarius* had highest abundance in decay
309 class III. All these genera have species known to form ectomycorrhizal symbiosis with
310 plant roots. Lindahl and Tunlid (2015) have demonstrated that some ectomycorrhizal
311 fungi have retained mechanisms of decomposition similar to those of brown- and white-
312 rot fungi and actively participate in the decomposition of organic matter. Nitrogen
313 availability is a major factor structuring ectomycorrhizal fungal communities. As
314 nitrogen increases, ectomycorrhizal fungal communities shift from taxa specialized in
315 N uptake under low-N conditions (e.g., *Cortinarius*, *Piloderma*) toward taxa
316 specialized in P uptake under high-N, low-P, acidified conditions (e.g., *Lactarius*)
317 (Lilleskov et al. 2002). The nutrient conditions of decaying logs and pH value of soil
318 under middle decomposition stage may be more suitable for the propagation of *Inocybe*,
319 *Russula* and *Lactarius*. Additionally, the abundance of *Fusarium* and *Humicola*
320 significantly increased during the decomposition process due to the strong positive
321 correlation between the organic matter decomposition and these two genera (Banerjee

322 et al. 2016), which can produce ligninase and cellulase to decompose refractory
323 substances.

324 **4.3 Response of soil fungal functional group to log decomposition**

325 According to the classification results of FUNGuild, the relative abundance of fungal
326 functional groups varied greatly with the trophic mode in the decomposing process of
327 logs. The soil underneath decaying Minjiang fir logs was dominated by saprotrophic
328 fungi. Wood saprotrophs and saprotroph were heterotrophic organisms that obtain a
329 major fraction of their metabolic carbon from dead organic matter (Lindahl and Tundia
330 2015). In this study, the relative abundance of wood saprotrophs and saprotroph
331 increased drastically at decay classes IV and V (Fig. 4), indicating that saprotrophic
332 fungi is favoured underneath the highly decayed logs. The relative abundance of
333 arbuscular mycorrhizal fungi increased during the decomposing process of logs.
334 Although the reason for the increase remains to be further studied, these arbuscular
335 mycorrhizal fungi likely play a role in early stages of decomposition. Besides,
336 ectomycorrhizal fungi in our study decreased during the decomposition process, which
337 might be related to the changes in soil nutrients and environmental
338 conditions. Additionally, the interaction between fungi with different trophic modes also
339 affected the proportion of fungal functional groups on guilds (Lindahl et al. 1999). The
340 results indicated that logs with different decay classes favoured different functional
341 groups on guilds of soil fungi.

342 **4.4 Influencing factors of soil fungal community composition**

343 Correlation between soil fungal community and physicochemical properties could be

344 attributed to the following reasons. Organic matter strongly affects the structure and
345 function of soil microbial communities (Grayston et al. 2004). Both carbon composition
346 and nitrogen addition can affect soil microbial community structure and function, such
347 as symbiotic mycorrhizal fungi (Morrison et al. 2016; Corrales et al. 2017; Zhang et al.
348 2020). Cao et al. (2016) have demonstrated that the concentration of phosphorus is
349 among the main factors influencing microbial activity at regional spatial scales. Besides,
350 water content was strongly correlated with the characteristics of the microbial
351 communities (Brockett et al. 2012). The relative abundance of some genera in our study
352 did not differ significantly among decay classes, suggesting that fungal genera in the
353 soil underneath decaying logs might respond to the changes in water content by the
354 allocation of limited resources. Soil pH is also a critical factor for the diversification of
355 fungi (Geisseler and Scow 2014). This is probably because pH indirectly affects the
356 structure of some microbial communities through interactions among soil elements (e.g.
357 precipitation of ions) (Lammel et al. 2018). The results showed that logs with different
358 decay classes provided various carbon and nutrient sources for soil fungal communities,
359 and highly decayed logs provided favorable water and pH conditions for functional
360 groups such as saprophytes.

361 **Conclusion**

362 This study examined the effects of decay classes of logs on the composition of soil
363 fungal communities in a subalpine forest. Our results showed that decaying logs altered
364 the compositions of soil fungal communities. The diversity of soil fungal community

365 and the relative abundance of dominant genera of soil differed significantly by the
366 decomposition of logs. Additionally, groups of fungi on guilds varied with decay classes
367 of logs, indicating that decaying logs favoured different functional groups of soil fungi
368 thus altering soil fungal community structures. Physicochemical properties (e.g.,
369 microbial biomass nitrogen, total phosphorus concentrations and water content) in our
370 study accounted to some extent for the difference of the fungal community structures
371 among decay classes. Therefore, the ecological effects of logs at differing decay classes
372 on soil fungal community were different. Log decomposition is a long-term and
373 complex biological process, so the response of soil microorganisms to the
374 decomposition of logs needs long-term experimental research.

375

376 **Author contributions**

377 W.Q.Y. and B.T. conceived and designed the experiment and contributed resources.
378 Q.W. completed laboratory analysis and led the writing of the manuscript. Q.W., B.T.,
379 Z.W., H.L., R.C., C.H.C., and Y.R.J. contributed to field work. W.Q.Y., and J.P.
380 provided writing assistance. All authors have read and agreed to the published version
381 of the manuscript.

382

383 **Funding:** This work was supported by the National Natural Science Foundation of
384 China (grant numbers 32071554, 31870602 and 31901295), and the Program of
385 Sichuan Excellent Youth Sci-Tech Foundation (grant numbers 2020JDJQ0052). We are
386 also grateful for the financial support from the China Scholarship Council, China

387 (202006910058), and the Spanish government project PID2019-110521GB-I00.

388

389 **Declaration of competing interest**

390 All authors certify that they have no affiliations with or involvement in any organization

391 or entity with any financial interest or non-financial interest in the subject matter or

392 materials discussed in this manuscript.

393 **Reference**

- 394 Altschul, S. 1990. Basic local alignment search tool. *Journal of Molecular Biology*. 215: 4
395 03-410. [http://doi.org/10.1016/S0022-2836\(05\)80360-2](http://doi.org/10.1016/S0022-2836(05)80360-2)
- 396 Asemaninejad, A., Thorn, R.G. and Branfireun, B.A. 2018. Climate change favours specific
397 fungal communities in boreal peatlands. *Soil Biology and Biochemistry*. 120: 28-3
398 6. <http://doi.org/10.1016/j.soilbio.2018.01.029>
- 399 Baldrian, P. and Banin, E. 2017. Forest microbiome: diversity, complexity and dynamics. *FEMS*
400 *Microbiology Reviews*. 41: 109-130. <http://doi.org/10.1093/femsre/fuw040>
- 401 Banerjee, S., Kirkby, C.A., Schmutter, D., Bissett, A., Kirkegaard, J.A. and Richardson, A.E. 2016.
402 Network analysis reveals functional redundancy and keystone taxa amongst bacterial and
403 fungal communities during organic matter decomposition in an arable soil. *Soil Biology*
404 *and Biochemistry*. 97: 188-198. <http://doi.org/10.1016/j.soilbio.2016.03.017>
- 405 Bani, A., Pioli, S., Ventura, M., Panzacchi, P., Borruso, L., Tognetti, R., Tonon, G. and Brusetti, L.
406 2018. The role of microbial community in the decomposition of leaf litter and deadwood.
407 *Applied Soil Ecology*. 126: 75-84. <http://doi.org/10.1016/j.apsoil.2018.02.017>
- 408 Barker, J.S. 2008. Decomposition of Douglas-fir coarse woody debris in response to differing
409 moisture content and initial heterotrophic colonization. *Forest Ecology and Management*.
410 255:598–604. <http://doi.org/10.1016/j.foreco.2007.09.029>
- 411 Błońska, E., Kacprzyk, M. and Spólnik, A. 2017. Effect of deadwood of different tree species in
412 various stages of decomposition on biochemical soil properties and carbon storage.
413 *Ecological Research*. 32: 193-203. <http://doi.org/10.1007/s11284-016-1430-3>
- 414 Bokulich, N.A., Subramanian, S., Faith, J.J., Gevers, D., Gordon, J.I., Knight, R., Mills, D.A. and
415 Caporaso, J.G. 2013. Quality-filtering vastly improves diversity estimates from Illumina
416 amplicon sequencing. *Nature Methods*. 10: 57-59. <http://doi.org/10.1038/nmeth.2276>
- 417 Brockett, B.F.T., Prescott, C.E. and Grayston, S.J. 2012. Soil moisture is the major factor
418 influencing microbial community structure and enzyme activities across seven bioge
419 oclimatic zones in western Canada. *Soil Biology and Biochemistry*. 44: 9-20. <http://doi.org/10.1016/j.soilbio.2011.09.003>
- 421 Brookes, P.C., Powlson, D.S. and Jenkinson, D.S. 1982. Measurement of microbial biomass
422 phosphorus in soil. *Soil Biology and Biochemistry*. 14: 319-329. [http://doi.org/10.1016/0038-0717\(82\)90001-3](http://doi.org/10.1016/0038-0717(82)90001-3)
- 424 Brookes, P.C., Landman, A., Pruden, G. and Jenkinson, D.S. 1985. Chloroform fumigation and the
425 release of soil nitrogen A rapid direct extraction method to measure microbial biomass
426 nitrogen in soil. *Soil Biology and Biochemistry*. 17: 837-842. [http://doi.org/10.1016/0038-0717\(85\)90144-0](http://doi.org/10.1016/0038-0717(85)90144-0)
- 428 Cao, H., Chen, R., Wang, L., Jiang, L., Yang, F., Zheng, S., Wang, G. and Lin, X. 2016. Soil pH,
429 total phosphorus, climate and distance are the major factors influencing microbial activity
430 at a regional spatial scale. *Scientific Reports*. 6: 25815. <http://doi.org/10.1038/srep25815>
- 431 Caporaso, J.G., Kuczynski, J., Stombaugh, J., Bittinger, K., Bushman, F.D., Costello, E.K.,
432 Fierer, N., Peña, A.G., Goodrich, J.K. and Gordon, J.I. 2010. QIIME allows analysis
433 of high-throughput community sequencing data. *Nature Methods*. 7: 335-336. <http://doi.org/10.1038/nmeth0510-335>
- 435 Chang, C., Wu, F., Wang, Z., Tan, B., Cao, R., Yang, W. and Cornelissen, J.H.C. 2019. E
436 ffects of epixylic vegetation removal on the dynamics of the microbial community

437 composition in decaying logs in an alpine forest. *Ecosystems*. 22: 1478-1496. [http://](http://doi.org/10.1007/s10021-019-00351-3)
438 [/doi.org/10.1007/s10021-019-00351-3](http://doi.org/10.1007/s10021-019-00351-3)

439 Chang, C., Wang, Z., Tan, B., Li, J., Cao, R., Wang, Q., Yang, W., Weedon, J.T. and Cornelissen,
440 J.H.C. 2020. Tissue type and location within forest together regulate decay trajectories of
441 *Abies faxoniana* logs at early and mid-decay stage. *Forest Ecology and Management*. 475:
442 118411. <http://doi.org/10.1016/j.foreco.2020.118411>

443 Chen, Y., Xu, T., Veresoglou, S.D., Hu, H., Hao, Z., Hu, Y., Liu, L., Deng, Y., Rillig, M.C. and Chen,
444 B. 2017a. Plant diversity represents the prevalent determinant of soil fungal community
445 structure across temperate grasslands in northern China. *Soil Biology and Biochemistry*.
446 110: 12-21. <http://doi.org/10.1016/j.soilbio.2017.02.015>

447 Chen, Y., Deng, Y., Ding, J., Hu, H., Xu, T., Li, F., Yang, G. and Yang, Y. 2017b. Distinct microbial
448 communities in the active and permafrost layers on the Tibetan Plateau. *Molecular Ecology*.
449 26: 6608-6620. <http://doi.org/10.1111/mec.14396>

450 Corrales, A., Turner, B.L., Tedersoo, L., Anslan, S. and Dalling, J.W. 2017. Nitrogen addition alters
451 ectomycorrhizal fungal communities and soil enzyme activities in a tropical montane forest.
452 *Fungal Ecology*. 27: 14-23. <http://doi.org/10.1016/j.funeco.2017.02.004>

453 Edgar, R.C. 2004. MUSCLE: multiple sequence alignment with high accuracy and high throughput.
454 *Nucleic Acids Research*. 32: 1792-1797. <http://doi.org/10.1093/nar/gkh340>

455 Edgar, R.C. 2010. Search and clustering orders of magnitude faster than blast. *Bioinformatics*. 26:
456 2460. <http://doi.org/10.1093/bioinformatics/btq461>

457 Edgar, R.C., Haas, B.J., Clemente, J.C., Quince, C. and Knight, R. 2011. UCHIME impro-
458 ves sensitivity and speed of chimera detection. *Bioinformatics*. 27: 2194-200. [http://](http://doi.org/10.1093/bioinformatics/btr381)
459 doi.org/10.1093/bioinformatics/btr381

460 Forrester, J.A., Mladenoff, D.J., Gower, S.T. and Stoffel, J.L. 2012. Interactions of tempera-
461 ture and moisture with respiration from coarse woody debris in experimental forest
462 canopy gaps. *Forest Ecology and Management*. 265: 124-132. [http://doi.org/10.101](http://doi.org/10.1016/j.foreco.2011.10.038)
463 [6/j.foreco.2011.10.038](http://doi.org/10.1016/j.foreco.2011.10.038)

464 Fukami, T., Dickie, I.A., Paula Wilkie, J., Paulus, B.C., Park, D., Roberts, A., Buchanan, P.K. and
465 Allen, R.B. 2010. Assembly history dictates ecosystem functioning: evidence from wood
466 decomposer communities. *Ecology Letters*. 13: 675-684. [http://doi.org/10.1111/j.1461-](http://doi.org/10.1111/j.1461-0248.2010.01465.x)
467 [0248.2010.01465.x](http://doi.org/10.1111/j.1461-0248.2010.01465.x)

468 Ganjegunte, G.K., Condrón, L.M., Clinton, P.W., Davis, M.R., Mahieu, N. 2004. Decomposition
469 and nutrient release from radiata pine (*Pinus radiata*) coarse woody debris. *For Ecology and*
470 *Management*. 187:197–211. [http://doi.org/10.1016/S0378-1127\(03\)00332-3](http://doi.org/10.1016/S0378-1127(03)00332-3)

471 Geisseler, D. and Scow, K.M. 2014. Long-term effects of mineral fertilizers on soil microo-
472 rganisms - A review. *Soil Biology and Biochemistry*. 75: 54-63. [http://doi.org/10.10](http://doi.org/10.1016/j.soilbio.2014.03.023)
473 [16/j.soilbio.2014.03.023](http://doi.org/10.1016/j.soilbio.2014.03.023)

474 Grayston, S.J., Campbell, C.D., Bardgett, R.D., Mawdsley, J.L., Clegg, C.D., Ritz, K., Griffiths,
475 B.S., Rodwell, J.S., Edwards, S.J., Davies, W.J., Elston, D.J. and Millard, P. 2004.
476 Assessing shifts in microbial community structure across a range of grasslands of differing
477 management intensity using CLPP, PLFA and community DNA techniques. *Applied Soil*
478 *Ecology*. 25: 63-84. [http://doi.org/10.1016/s0929-1393\(03\)00098-2](http://doi.org/10.1016/s0929-1393(03)00098-2)

479 Harmon, M.E., Franklin, J.F., Swanson, F.J., Sollins, P. and Gregory, S.V. 1986. Ecology of coarse
480 woody debris in temperate ecosystems. *Advances in Ecological Research*. 15: 133-302.

481 [http://doi.org/10.1016/S0065-2504\(03\)34002-4](http://doi.org/10.1016/S0065-2504(03)34002-4).

482 Jarosław, L., Ewa, B., Wojciech, P. and Małgorzata, W. 2017. How the deadwood of different tree
483 species in various stages of decomposition affected nutrient dynamics? *Journal of Soils and*
484 *Sediments*. 18: 2759–2769. <http://doi.org/10.1007/s11368-017-1858-2>

485 Kõljalg, U., Nilsson, R.H., Abarenkov, K., Tedersoo, L., Taylor, A.F., Bahram, M., Bates, S.T.,
486 Bruns, T.D., Bengtsson-Palme, J., Callaghan, T.M., Douglas, B., Drenkhan, T., Eberhardt,
487 U., Duenas, M., Grebenc, T., Griffith, G.W., Hartmann, M., Kirk, P.M., Kohout, P., Larsson,
488 E., Lindahl, B.D., Lucking, R., Martin, M.P., Matheny, P.B., Nguyen, N.H., Niskanen, T.,
489 Oja, J., Peay, K.G., Peintner, U., Peterson, M., Poldmaa, K., Saag, L., Saar, I., Schussler,
490 A., Scott, J.A., Senes, C., Smith, M.E., Suija, A., Taylor, D.L., Telleria, M.T., Weiss, M.
491 and Larsson, K.H. 2013. Towards a unified paradigm for sequence-based identification of
492 fungi. *Molecular Ecology*, 22: 5271-7. <http://doi.org/10.1111/mec.12481>

493 Lammel, D.R., Barth, G., Ovaskainen, O., Cruz, L.M., Zanatta, J.A., Ryo, M., de Souza,
494 E.M. and Pedrosa, F.O. 2018. Direct and indirect effects of a pH gradient bring in
495 sights into the mechanisms driving prokaryotic community structures. *Microbiome*.
496 61: 106. <http://doi.org/10.1186/s40168-018-0482-8>

497 Lankau, R.A. and Keymer, D.P. 2016. Ectomycorrhizal fungal richness declines towards the host
498 species' range edge. *Molecular Ecology*. 25: 3224-3241. <http://doi.org/10.1111/mec.13628>

499 Lassauce, A., Paillet, Y., Jactel, H. and Bouget, C. 2011. Deadwood as a surrogate for
500 est biodiversity: Meta-analysis of correlations between deadwood volume and speci
501 es richness of saproxylic organisms. *Ecological Indicators*. 11: 1027-1039. <http://doi.org/10.1016/j.ecolind.2011.02.004>

502

503 Leopold, D.R., Wilkie, J.P., Dickie, I.A., Allen, R.B., Buchanan, P.K. and Fukami, T. 201
504 7. Priority effects are interactively regulated by top-down and bottom-up forces: ev
505 idence from wood decomposer communities. *Ecology Letters*. 20: 1054-1063. <http://doi.org/10.1111/ele.12803>

506

507 Lilleskov, E.A., Fahey, T.J., Horton, T.R. and Lovett, G.M. 2002. Belowground ectomycorrhizal
508 fungal community change over a nitrogen deposition gradient in Alaska. *Ecology*. 83: 104-
509 115. <http://doi.org/10.2307/2680124>

510 Lindahl, B., Stenlid, J., Olsson, S. and Finlay, R. 1999. Translocation of ³²P between interacting
511 mycelia of a wood- decomposing fungus and ectomycorrhizal fungi in microcosm systems.
512 *New Phytologist*. 144: 183-193. <http://doi.org/10.2307/2588287>

513 Lindahl, B.D., Ihrmark, K., Boberg, J., Trumbore, S.E., Högborg, P., Stenlid, J. and Finlay, R.D.
514 2007. Spatial separation of litter decomposition and mycorrhizal nitrogen uptake in a boreal
515 forest. *New Phytologist*. 173: 611-620. <http://doi.org/10.1111/j.1469-8137.2006.01936.x>

516 Lindahl, B.D. and Tunlid, A. 2015. Ectomycorrhizal fungi - potential organic matter decomposers,
517 yet not saprotrophs. *New Phytologist*. 205: 1443-1447. <http://doi.org/10.1111/nph.13201>

518 Lu, R. 1999. *Soil and Agro-chemical Chemistry Analytical Methods*. China Agricultural Science
519 and Technology Press, Beijing.

520 Magoč, T. and Salzberg, S.L. 2011. FLASH: fast length adjustment of short reads to impr
521 ove genome assemblies. *Bioinformatics*. 27: 2957-2963. <http://doi.org/10.1093/bioinformatics/btr507>

522

523 Mäkipää, R., Rajala, T., Schigel, D., Rinne, K.T., Pennanen, T., Abrego, N. and Ovaskaine
524 n, O. 2017. Interactions between soil- and dead wood-inhabiting fungal communitie

525 s during the decay of Norway spruce logs. *ISME Journal*. 11: 1964-1974. <http://doi.org/10.1038/ismej.2017.57>

526

527 Minnich, C., PerÅ;oh, D., Poll, C. and Borken, W. 2021. Changes in chemical and microbial soil
528 parameters following 8 years of deadwood decay: an experiment with logs of 13 tree
529 species in 30 forests. *Ecosystems*. 24: 955-967. [http://doi.org/10.1007/s10021-020-00562-](http://doi.org/10.1007/s10021-020-00562-z)
530 [z](http://doi.org/10.1007/s10021-020-00562-z)

531 Moorhead, D.L., Lashermes, G. and Sinsabaugh, R.L. 2012. A theoretical model of C- and N-
532 acquiring exoenzyme activities, which balances microbial demands during decomposition.
533 *Soil Biology and Biochemistry*. 53: 133-141. <http://doi.org/10.1016/j.soilbio.2012.05.011>

534 Morrison, E.W., Frey, S.D., Sadowsky, J.J., van Diepen, L.T.A., Thomas, W.K. and Pringle,
535 A. 2016. Chronic nitrogen additions fundamentally restructure the soil fungal com-
536 munity in a temperate forest. *Fungal Ecology*. 23: 48-57. <http://doi.org/10.1016/j.funeco.2016.05.011>

537

538 Nguyen, N.H., Song, Z., Bates, S.T., Branco, S., Tedersoo, L., Menke, J., Schilling, J.S. a
539 nd Kennedy, P.G. 2016. FUNGuild: An open annotation tool for parsing fungal co-
540 mmunity datasets by ecological guild. *Fungal Ecology*. 20: 241-248. <http://doi.org/10.1016/j.funeco.2015.06.006>

541

542 Oksanen, J., Blanchet, F.G., Kindt, R., Legendre, P., Minchin, P.R., O'Hara, R.B., Simpson, G.L.,
543 Solymos, P., Stevens, M.H.H. and Wagner, H.H. 2013. *Vegan: community ecology package*.
544 R package version. 2.0-10. <http://CRAN.R-project.org/package=vegan>.

545 R Development Core Team. 2015. *R: A Language and Environment for Statistical Computing*. R
546 Foundation for Statistical Computing, Vienna. <http://www.r-project.org/>

547 Semchenko, M., Leff, J.W., Lozano, Y.M., Saar, S., Davison, J., Wilkinson, A., Jackson, B.
548 G., Pritchard, W.J., De Long, J.R., Oakley, S., Mason, K.E., Ostle, N.J., Baggs, E.
549 M., Johnson, D., Fierer, N. and Bardgett, R.D. 2018. Fungal diversity regulates pla-
550 nt-soil feedbacks in temperate grassland. *Science Advances*. 4: eaau4578. <http://doi.org/10.1126/sciadv.aau4578>

551

552 Tedersoo, L., Bahram, M., Põlme, S., Kõljalg, U., Yorou, N.S., Wijesundera, R., Ruiz, L.V., Vasco-
553 Palacios, A.M., Thu, P.Q., Suija, A., Smith, M.E., Sharp, C., Saluveer, E., Saitta, A., Rosas,
554 M., Riit, T., Ratkowsky, D., Pritsch, K., Põldmaa, K., Piepenbring, M., Phosri, C., Peterson,
555 M., Parts, K., Pärtel, K., Otsing, E., Nouhra, E., Njouonkou, A.L., Nilsson, R.H., Morgado,
556 L.N., Mayor, J., May, T.W., Majuakim, L., Lodge, D., Lee, S.S., Larsson, K.-H., Kohout,
557 P., Hosaka, K., Hiiesalu, I., Henkel, T.W., Harend, H., Guo, L.D., Greslebin, A., Grelet, G.,
558 Geml, J., Gates, G., Dunstan, W., Dunk, C., Drenkhan, R., Dearnaley, J., Kesel, A.D., Dang,
559 T., Chen, X., Buegger, F., Brearley, F.Q., Bonito, G., Anslan, S., Abell, S. and Abarenkov,
560 K. 2014. Fungal biogeography. Global diversity and geography of soil fungi. *Science*. 346:
561 6213. <http://doi.org/10.1126/science.1256688>

562 Vance, E.D., Brooke, P.C. and Jenkinson, D.S. 1987. An extraction method for measuring
563 soil microbial biomass C. *Soil Biology and Biochemistry*. 19: 703-707. [http://doi.or](http://doi.org/10.1016/0038-0717(87)90052-6)
564 [g/10.1016/0038-0717\(87\)90052-6](http://doi.org/10.1016/0038-0717(87)90052-6)

565 Wang, Q., Yang, W., Li, H., Wang, Z., Chang, C., Cao, R. and Tan, B. 2021. Changes in carbon,
566 nitrogen and phosphorus stoichiometry in decaying logs with gap positions in a subalpine
567 forest. *Journal of Plant Ecology*. 14: 692-701. <http://doi.org/10.1093/jpe/rtab023>

568 Wojciech, P., Ewa, B., Jarosław, L. and Martin, L. 2019. A comparison of C:N:P stoichio

569 metry in soil and deadwood at an advanced decomposition stage. *Catena*. 179: 1-5.
570 <http://doi.org/10.1016/j.catena.2019.03.025>

571 Xiao, S., Wu, F., Yang, W., Chang, C., Li, J. and Wang, B. 2014. Understory biomass and its
572 characteristics as affected by forest gap in the alpine forest ecosystem in West Sichuan.
573 *Ecology and Environmental Sciences*. 23: 1515-1519. <http://doi.org/10.3969/j.issn.1674-5906.2014.09.018>
574

575 Xiao, S., Wu, F., Yang, W., Chang, C., Li, J., Wang, B. and Cao, Y. 2016. Woody debris storage and
576 its distribution in a dark coniferous forest in the alpine-gorge area. *Acta Ecologica Sinica*.
577 36: 1352–1359. <http://doi.org/10.5846/stxb201407171458>

578 Yang, W. and Wu, F. 2021. The ecosystem processes and adaptive management of the subalpine
579 coniferous forests in the upper reaches of Yangtze river. Beijing: Science Press.

580 Zalamea, M., González, G., Ping, C. and Michaelson, G. 2007. Soil organic matter dynamics under
581 decaying wood in a subtropical wet forest: effect of tree species and decay stage. *Plant and*
582 *Soil*. 296: 173-185. <http://doi.org/10.1007/s11104-007-9307-4>.

583 Zalamea, M., González, G. and Lodge, D. 2016. Physical, chemical, and biological propert
584 ies of soil under decaying wood in a tropical wet forest in Puerto Rico. *Forests*.
585 7: 168-186. <http://doi.org/10.3390/f7080168>

586 Zhang, J., Kobert, K., Flouri, T. and Stamatakis, A. 2014. PEAR: a fast and accurate Illu
587 mina Paired-End reAd mergeR. *Bioinformatics*. 30: 614-20. <http://doi.org/10.1093/bioinformatics/btt593>
588

589 Zhang, L., Ren, Y., Yang, K., Li, Z., Tan, B., Liu, Y., Li, H., You, C., Liu, S., Wang, L.,
590 Yin, R., Zhang, J. and Xu, Z. 2021. Immediate and legacy effects of snow exclu
591 sion on soil fungal diversity and community composition. *Forest Ecosystems*. 8: 2
592 2. <http://doi.org/10.1186/s40663-021-00299-8>

593 Zhang, X., Chen, X., Liu, M., Xu, Z. and Wei, H. 2020. Coupled changes in soil organic carbon
594 fractions and microbial community composition in urban and suburban forests. *Scientific*
595 *Reports*. 10: 15933. <http://doi.org/10.1038/s41598-020-73119-8>
596

597 **Figure captions**

598 **Fig. 1** Relative abundances (>1%) of soil fungi at the phylum (a) and genus (b) levels in the
599 control (CK) and the decay classes (ranked by the sum of means). Groups accounting for <1% of
600 the relative abundance are integrated into ‘others’. CK, control; I - V, decay classes I -V. Data are
601 the means of five samples for each group.

602

603 **Fig. 2** Nonmetric multidimensional scaling (NMDS) analysis of the composition of soil fungal
604 communities based on Bray-Curtis distances. CK, control; I - V, decay classes I -V.

605

606 **Fig. 3** Fungal functional groups (guild) under different treatments (mean \pm *SD*, $n = 5$). CK,
607 control; I - V, decay classes I -V. Different letters indicate significant differences ($P < 0.05$) among
608 the decay classes.

609

610 **Fig. 4** Heatmap of the relative abundance of dominant genera (a) and the proportion of functional
611 groups (b) from soil fungi. Blue denotes a low relative value across a taxon (row); red denotes a
612 high relative value. The color key for the Z score indicates correspondence between the blue-red
613 coloring and standard deviations from the mean value of each taxon. CK, control; I-V, decay
614 classes I-V.

615

616 **Fig. 5** Correlation heatmap of soil physicochemical properties and the relative abundances of soil
617 fungi at phylum (a) and genus (b) level. MBC: microbial biomass carbon, MBN: microbial biomass
618 nitrogen, MBP: microbial biomass phosphorus, OC: organic carbon, TN: total nitrogen, TP: total
619 phosphorus, WC: water content. * $P < 0.05$, ** $P < 0.01$.

620

621 **Table 1** Number of operational taxonomic units (OTUs) and indices of Shannon and Simpson
 622 diversity for the fungal communities in the soil underneath decaying logs for the decay classes (mean
 623 $\pm SD$, $n = 5$). CK, control; I - V, decay classes I -V. Different letters within a column indicate
 624 significant differences ($P < 0.05$) among the classes.
 625

Class	No. OTUs	Shannon	Simpson
CK	607 \pm 199a	3.84 \pm 0.72b	0.734 \pm 0.055c
I	684 \pm 48a	4.63 \pm 0.40b	0.832 \pm 0.050b
II	663 \pm 166a	3.76 \pm 0.48b	0.722 \pm 0.053c
III	845 \pm 160a	6.14 \pm 0.45a	0.951 \pm 0.016a
IV	790 \pm 178a	6.07 \pm 0.92a	0.954 \pm 0.029a
V	890 \pm 115a	5.94 \pm 1.09a	0.929 \pm 0.059a

626